

Manuscript Details

Manuscript number	APAC_2019_446_R1
Title	Microphone position and noise exposure assessment of building façades
Article type	Research Paper

Abstract

The prevention or reduction of the exposure of the population to environmental noise is the fundamental objective of the European Noise Directive (END). To this end, strategic noise maps are considered as the basic tool and on-site measurements play an important role in its successful implementation. In this regard, the ISO 1996 standards are the reference for the measurement and assessment of environmental noise, but their application may be complex in many cases. It is therefore necessary to find urban scenarios in which the effects of the placement of the measuring equipment with respect to the façade on noise exposure levels can be analysed. In this study, an educational building was selected for an analysis of the differences in the weekly sound level from road traffic between a microphone flush-mounted on a plate at the façade and another placed between 0.5 and 2.0 m from it. The recommendations in Annex B of ISO 1996-2 were followed in the placement of the microphones. A broadband analysis shows that similar results were found for the four distances analysed, but that variations of up to 0.6 dBA above the reference value arose. An analogous study using frequency octave bands shows differences higher than 2 dB between the measured configurations for bands under 250 Hz. Based on the distance range given in ISO 1996-2 for the position of a microphone in front of a reflecting surface, the results suggest that the most appropriate option for accurately assessing the sound level incident on the façade of buildings is to place the microphone at a distance of 2 m if the guidelines of the ISO 1996-2 standard can be met.

Keywords	environmental pollution; ISO 1996; measurement uncertainty; reflecting surface; health; road traffic noise
Taxonomy	Environmental Acoustics, Outdoor Propagation of Sound, Building Acoustics
Manuscript category	Europe and Rest of the World
Corresponding Author	Guillermo Rey Gozalo
Corresponding Author's Institution	Universidad de Extremadura
Order of Authors	David Montes González, Juan Miguel Barrigón, Guillermo Rey Gozalo, Pedro Atanasio Moraga
Suggested reviewers	Gianluca Memoli, Mário Mateus, Aleksandras Jagniatinskis

Submission Files Included in this PDF

File Name [File Type]

Cover Letter.docx [Cover Letter]

response_to_reviewers.docx [Response to Reviewers]

Manuscript.docx [Manuscript File]

figure_1.tif [Figure]

figure_2a.tif [Figure]

figure_2b.tif [Figure]

figure_3.tif [Figure]

figure_4a.tif [Figure]

figure_4b.tif [Figure]

figure_5.tif [Figure]

figure_6.tif [Figure]

figure_7.tif [Figure]

declaration-of-competing-interests.docx [Conflict of Interest]

agreement_authors.docx [Author Agreement]

Manuscript_revised with changes marked.docx [Supporting File]

To view all the submission files, including those not included in the PDF, click on the manuscript title on your EVISE Homepage, then click 'Download zip file'.

23 road traffic between a microphone flush-mounted on a plate at the façade and another
24 placed between 0.5 and 2.0 m from it. The recommendations in Annex B of ISO 1996-2
25 were followed in the placement of the microphones. A broadband analysis shows that
26 similar results were found for the four distances analysed, but that variations of up to
27 0.6 dBA above the reference value arose. An analogous study using frequency octave
28 bands shows differences higher than 2 dB between the measured configurations for
29 bands under 250 Hz. Based on the distance range given in ISO 1996-2 for the position
30 of a microphone in front of a reflecting surface, the results suggest that the most
31 appropriate option for accurately assessing the sound level incident on the façade of
32 buildings is to place the microphone at a distance of 2 m if the guidelines of the ISO
33 1996-2 standard can be met.

34 **Keywords:** environmental pollution; ISO 1996; measurement uncertainty; reflecting
35 surface; health; road traffic noise.

36 **1. INTRODUCTION**

37 *In situ* noise measurements are taken into consideration in the European Noise
38 Directive (END) [1] for the elaboration of strategic noise maps in relation to the
39 exposure of the population to environmental noise. They are also recognised as essential
40 by the Good Practice Guide [2] in the development and validation of computational
41 methods for noise mapping. This guide states that measurements have many additional
42 functions to fulfil in the successful application of the European Noise Directive (END)
43 [1], and especially in the development of action plans and verification of the real effects
44 of actions that are carried out [3-5].

45 Annex II of END [1] and its subsequent modification [6] propose the use of the
46 series of ISO 1996 standards [7,8] for the measurement and assessment of

47 environmental noise; these have recently been updated, and many works use them in the
48 planning and development of outdoor noise studies [9-13].

49 The ISO 1996-2 standard [8] proposes certain corrections to be applied to
50 measured sound levels to consider the effects of sound reflections on building façades
51 [14]. Depending on the position of the microphone, three different scenarios are
52 identified. Firstly, in the case of a microphone located at the façade of a building, an
53 interesting issue is introduced in the new version of the standard [8]. The correction
54 required to obtain a free field is 5.7 dBA if the conditions in Annex B are met;
55 otherwise, this figure can reach 6 dBA. Secondly, if the microphone is between 0.5 and
56 2.0 m in front of the surface, the value of this correction is 3 dB. Finally, for the free-
57 field position (the reference condition), no correction is applied.

58 The new version of the ISO 1996-2 standard [8] includes a new aspect with
59 respect to the previous version [15] in relation to the calculation of the uncertainty
60 associated with the selection of microphone location. A table is presented in Annex B
61 that gives the uncertainty standard for corrections to the reflections of different
62 microphone locations relative to vertical reflecting surfaces for a road traffic sound
63 source. This topic has been the subject of study in many papers where road traffic is
64 considered the main source [16-21], and also in cases where sound impacts on the
65 façades of buildings at very different angles, for example in the case of aircraft noise
66 [22]. Hall et al. [16], Quirt [17] and Jagniatinskis et al. [18] conclude that a value of 3
67 dB for the correction is usually suitable when the microphone is 2.0 m from the façade.
68 Memoli et al. [19] reach a similar conclusion for microphone positions of between 0.5
69 m and 2.0 m from the surface, but in this case, deviations from the proposed value of up
70 to about 1 dB are detected. Montes et al. [20] report several effects on corrections
71 associated with parking lines. Finally, Mateus et al. [21] describe the possible influence

72 of wind on the registered values for corrections in a case where the microphones were
73 located at about 150 m from the sound source.

74 Some of these conditions have been analysed in previous studies [23,24], since
75 this microphone location is frequently used in research about environmental noise
76 [25,26] and in *in situ* measurements based on the ISO 11819-4 standard [27] for a
77 modified version of the statistical pass-by method or the ISO 10847 standard [28] for
78 the determination of insertion loss of outdoor noise barriers. Barrigón et al. [23]
79 highlight disparities in the values obtained in this measurement configuration of up to 2
80 dB from the correction of 6 dB established by the previous version of the ISO 1996-2
81 [15] standard. In the same vein, Montes et al. [24] report a study in which the
82 microphone was placed on a reflective surface using plates of different materials. The
83 experimental values of the corrections varied at the different measurement points and
84 were lower than 6 dB, reaching average values of up to 5.2 dB in one of the studied
85 cases.

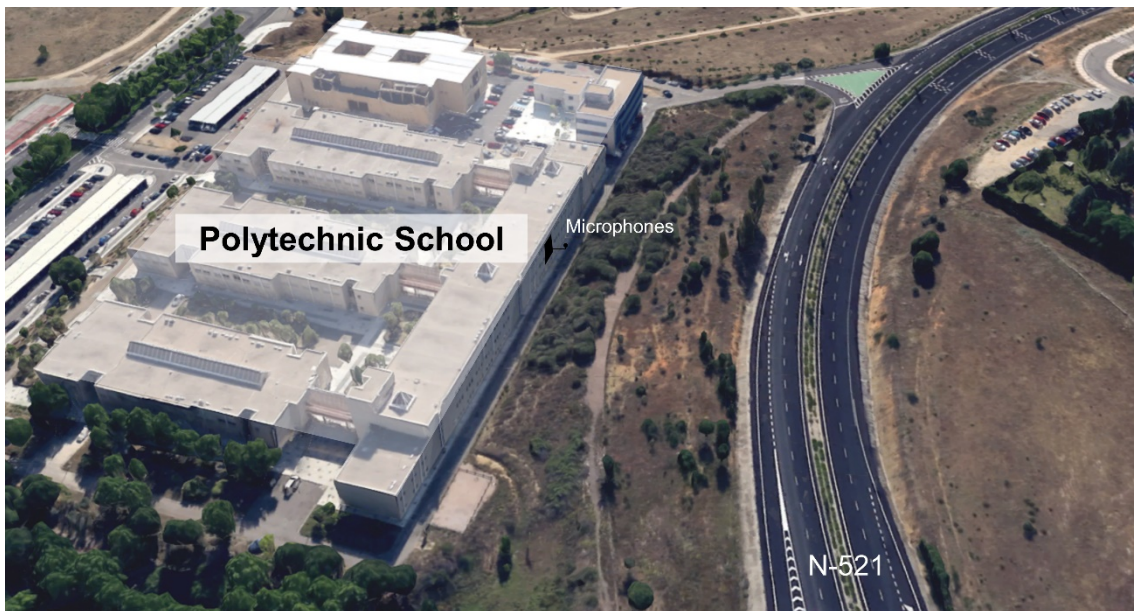
86 A-weighting is generally used in the assessment of the impact of environmental
87 noise on the population. This means that low frequencies have a lower weight in the
88 final value of the sound indicators than medium or high frequencies. However, it is well
89 known that the behaviour of sound waves is highly dependent on the frequency,
90 meaning that each octave band may show different behaviour depending on the location
91 of the microphone. The use of A-weighting to assess annoyance in environmental noise
92 has been analysed and reviewed in the literature. For example, in a study conducted by
93 Schomer et al. [29], the loudness-level-weighted equivalent level was proposed for
94 assessment of the annoyance from environmental noise. The results indicate that
95 compared with the A-weighting, a loudness-level weighting can order and assess
96 transportation noise sources more effectively. Nilsson et al. report that with similar A-

97 weighted sound pressure levels, sounds with higher low-frequency levels were louder
98 and more annoying than sounds with lower levels [30], or suggest that an evaluation of
99 the annoyance-reduction efficiency of noise barriers should not only rely on the
100 expected reduction in A-weighted sound pressure level but should also take into account
101 the associated increase in the relative level of low-frequency sound [31]. Even in studies
102 conducted in urban green areas, the linearly weighted equivalent level showed a better
103 relationship with traffic noise annoyance than an A-weighted value [10]. The effects of
104 low-frequency noise on humans have also been studied by various researchers. A
105 review of low-frequency noise sources and their effects can be found in the work of
106 Berglund et al. [32]. Another study shows that the prevalence of annoyance, disturbed
107 concentration and rest was greater in areas with low-frequency noise exposure, and that
108 the A weighting was a poor predictor of annoyance due to low-frequency noise [33].

109 Taking into account these aspects of the new version of ISO 1996-2 [8], the
110 importance of the frequency weighting used to determine environmental noise levels
111 and the role of *in situ* measurements in the assessment and management of
112 environmental noise [1], this paper presents a broad study of the corrections proposed
113 by this standard by means of one-week measurements for different positions of the
114 microphone. A broadband analysis is first carried out in which the weekly measured
115 values are analysed globally. The distribution of the correction values is then analysed
116 to determine their variability according to the time of day and the measured sound level.
117 Finally, a study in octave frequency bands of the corrections is presented in order to
118 deepen the experimental knowledge of sound reflections and interference phenomena
119 that take place in front of vertical surfaces in urban environments.

120 **2. METHODOLOGY**

121 This study was carried out at the Polytechnic School of the University of
122 Extremadura (Cáceres, Spain). The measurement devices were located on the façade of
123 a teaching area in the Common Building of the Polytechnic School, which has direct
124 view of the N-521 road (Fig. 1). This is an acoustically sensitive building with
125 educational use and is directly affected by one of the main access roads to the city [34-
126 36].



127

128

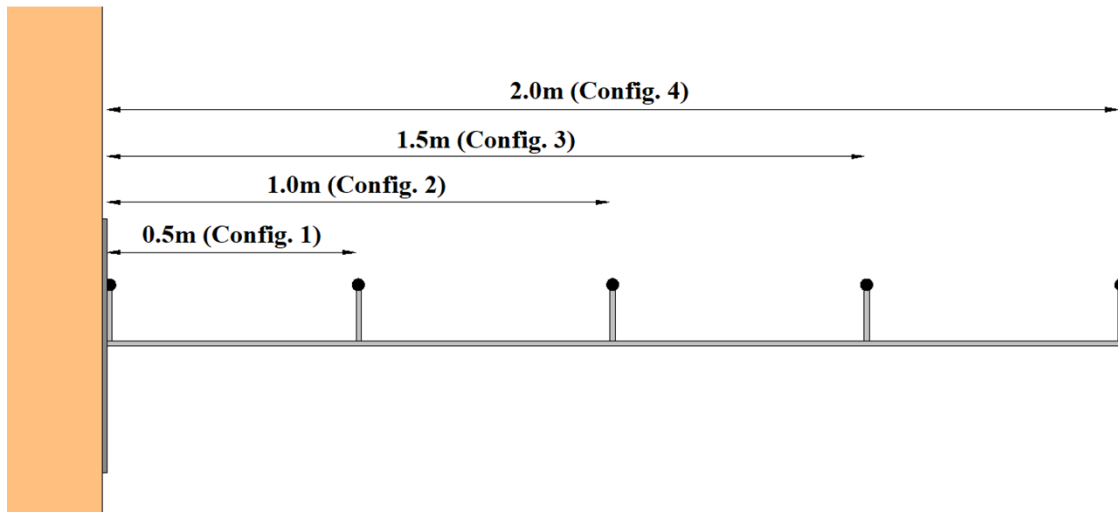
Fig. 1. Polytechnic School (Google Earth)

129 Two class 1 sound level meters (Brüel & Kjær type 2250 Light) were used to
130 establish four different measurement configurations. In each of these configurations, the
131 microphone of one of the sound level meters (façade microphone) was placed flush with
132 the façade of the building on a 6-mm-thick metal plate with effective dimensions of
133 0.535 x 0.705 m. The microphone of the second sound level meter (mobile microphone)
134 was placed at varying distances (0.5, 1.0, 1.5 and 2.0 m) from the façade (Fig. 2a). Both
135 microphones were placed at a height of 7.75 m above the ground. Fig. 2b shows a

136 photograph of configuration 4, in which one of the microphones is placed on the façade
137 and the other one is 2.0 m away from it.

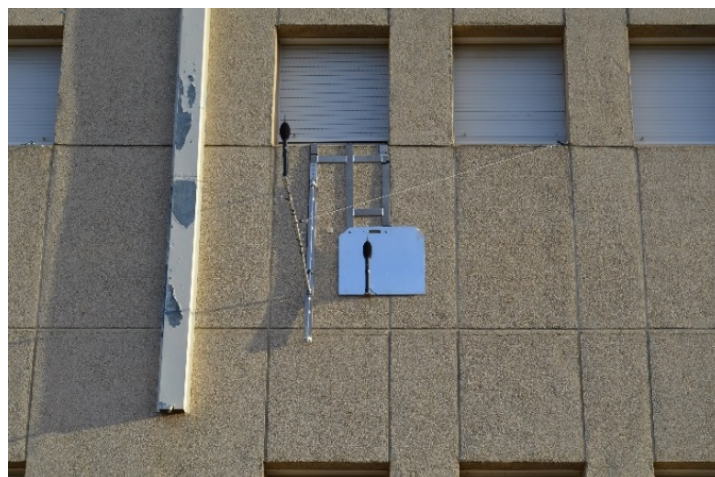
138

139 a)



140

141 b)



142

143 Fig. 2. (a) Diagram of the four configurations; and (b) photograph of configuration 4

144 Measurements of the sound levels were carried out with an integration time of one
145 minute over the period of a week in each configuration. In this way, an analysis of the

146 results for different integration time intervals was carried out. A detailed analysis of the
147 time of the day at which each measurement was made and the values of the registered
148 sound levels was also performed. Finally, a spectral analysis of the differences between
149 the two microphones in each of the four configurations was performed. To this end, the
150 octave bands included in the general frequency range defined by the ISO 1996-2
151 standard were considered.

152 The guidelines provided by the ISO 1996-2 standard in its normative part and
153 Annex B [8] were used as a basis for placing the microphones: the characteristics of the
154 façade and the measurement devices, the positions of the microphones and distances to
155 the edges of the reflective surface, and the relationship of the distances between façade,
156 microphone and sound source.

157 In addition to the sound levels registered, the meteorological conditions were
158 monitored at one minute intervals to verify that the values were within the ranges
159 recommended for the International Standards and thus discard their influence on the
160 variability of the sound levels. The ANSI S12.18 [37] and ISO 1996-2 [8] guidelines
161 were considered in the weather station location.

162 **3. RESULTS AND DISCUSSION**

163 This section presents an analysis and discussion of the results of measurements
164 carried out using the procedure described in Section 2. A study of the broadband results
165 is first presented in Section 3.1, in accordance with ISO 1996-2. A spectral analysis in
166 octave bands is then carried out in Section 3.2.

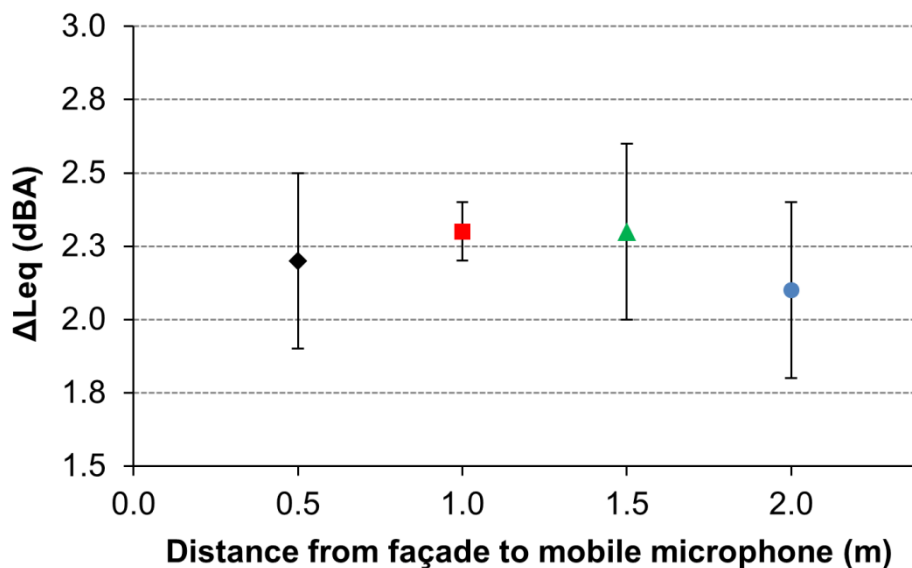
167 **3.1. Broadband results**

168 *3.1.1. Global differences*

169 As previously indicated, four measurement configurations were analysed. In each
 170 of these configurations, one of the microphones was placed on the reflective surface
 171 (façade microphone) and the other at varying distances from the façade (mobile
 172 microphone): 0.5 m (configuration 1), 1.0 m (configuration 2), 1.5 m (configuration 3)
 173 and 2.0 m (configuration 4). Using measurements taken over one week, the difference in
 174 the equivalent sound level registered by façade microphone ($LAeq_{façade}$) and mobile
 175 microphone ($LAeq_{mobile}$) for an integration time of one hour was calculated for each
 176 configuration i (Eq. 1)

$$177 \quad \Delta LAeq_i = LAeq_{façade} - LAeq_{mobile} \quad (1)$$

178 Fig. 3 shows the average and standard deviation of the sound level differences
 179 between the two microphones obtained in each measurement configuration. It can be
 180 seen in Fig 3 that the interval of variability in configuration 2 is smaller than in the rest.
 181 This difference in the variability interval may be mainly related to a greater uniformity
 182 in the equivalent sound level at nights, that do not reach values as low as the other
 183 monitoring weeks, and a lower occurrence of sound events other than traffic noise (main
 184 sound source).



185

186 Fig. 3. Average and standard deviation of the differences in sound levels recorded
187 by both microphones for each measurement configuration

188 As mentioned in the introduction, the ISO 1996-2 standard introduces certain
189 corrections with respect to the free-field situation: 3 dBA for a microphone positioned at
190 a distance of between 0.5 and 2.0 m from the façade, and 5.7 dBA for a microphone at a
191 reflective surface (if the conditions of Annex B are met [23]). Therefore, since these
192 conditions are fulfilled at this measurement point, the expected value for the difference
193 in the equivalent sound level between the two microphone positions is 2.7 dBA (Eq. 1).
194 Taking into account the average values shown in Fig. 3 for all the measurement
195 configurations, variations of up to 0.6 dB are observed relative to the expected value. In
196 view of the greatest value of 0.5 dB for the range indicated in Annex B of ISO 1996-2
197 for the standard uncertainty of corrections to the reflections of different microphone
198 locations relative to vertical reflecting surfaces for road traffic noise incident from all
199 angles, the average values of the differences between the equivalent sound levels
200 obtained would be located at the lower limit, or even outside the specified range. From
201 the standard deviations shown in Fig. 3, it can be seen that: a) the lower values are
202 clearly below the expected range in three of the four measurement configurations; and
203 b) the value of 2.7 dB is outside the range of variability found in any measurement
204 configuration. However, if all sources of uncertainty included in the ISO 1996-2
205 standard (measuring system, meteorological conditions, etc.) were considered, the
206 measurement uncertainty would be at least 1 dB. In any case, this variation with respect
207 to the ISO 1996-2 standard indicates the need for future studies. It is important that
208 studies are carried out using long-term measures and according to the requirements
209 indicated in the standard. This would allow to compare results obtained at different

210 sites, in order to conclude whether this bias is a feature of the site or of the ISO 1996-2
 211 methodology.

212 An inferential statistical analysis comparing the average values between the
 213 different measurement configurations shows that the average differences found in the
 214 2.0 m configuration (configuration 4) are statistically lower than in the other
 215 configurations (see Table 1).

216

217 Table 1. *P*-values for pairwise comparisons of configurations using a *t*-test for ΔL_{eq}
 218 (dBA)

	Configuration	1	2	3
ΔL_{eq} (dBA)	2	> 0.05	-	-
	3	> 0.05	> 0.05	-
	4	< 0.01	< 0.001	< 0.001

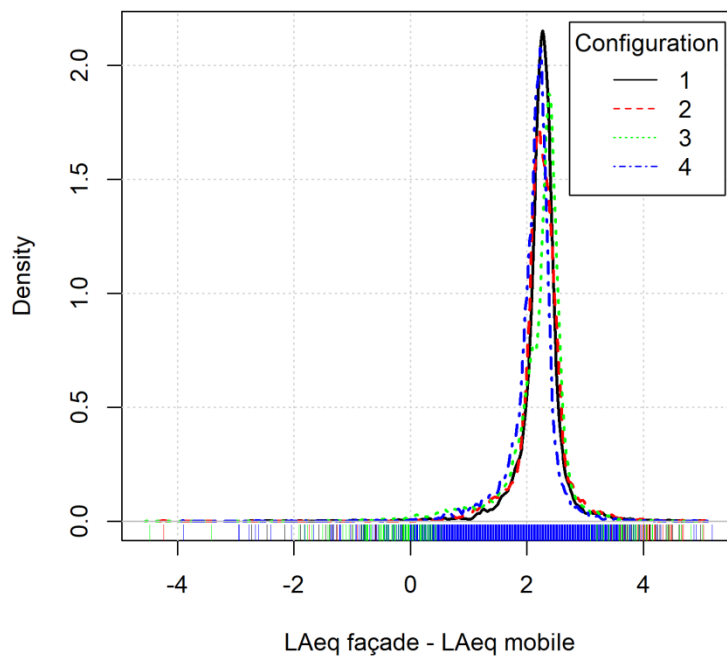
219

220 *3.1.2. Distribution of sound differences*

221 The average weekly sound levels were considered in the analysis carried out in the
 222 previous section, but these corrections may vary depending on the values of the sound
 223 levels and their temporality [38-45]. This aspect may be of interest in relation to the
 224 research line focused on dynamic noise maps [46-47]. A study of the statistical
 225 distribution of the differences in the sound levels measured by both microphones was
 226 carried out in order to analyse these possibilities, using an integration time of one
 227 minute, as used in the measurements. This allows to carry out a detailed analysis
 228 relative to both the time and the measured levels. It is therefore possible to find a
 229 structure related to the times at which the measurements were made or the values of the
 230 sound levels when the differences are analysed for a short integration time.

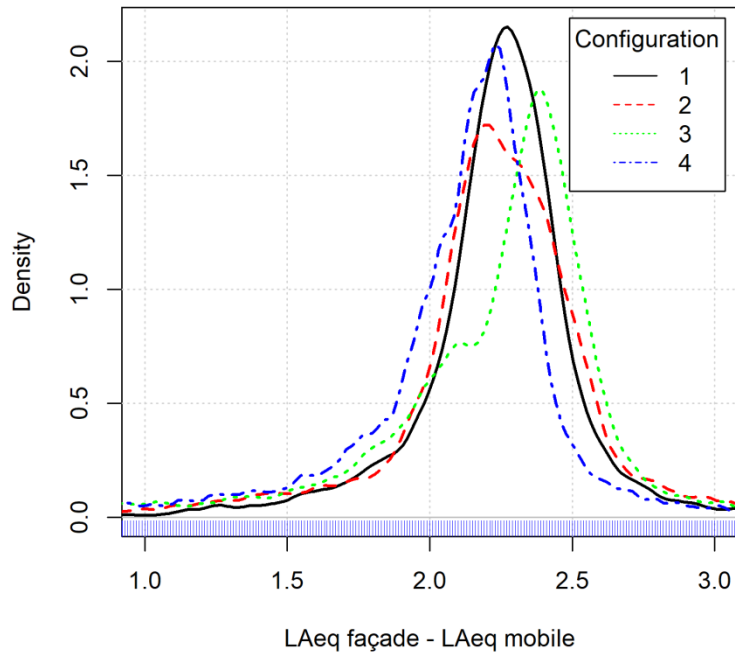
231 The distribution of the differences in sound levels was first analysed for each
232 measurement configuration through the density function (Fig. 4). The values of the
233 differences in sound levels are represented on the abscissa in Fig. 4, while the ratio of
234 the values per unit variable (density) is on the ordinate. Therefore, if in a very small
235 interval on the x-axis, for example, in a quarter of a dBA, there is more than a quarter of
236 the data, more than 100% of the data per dBA unit is being reached in that area. When
237 this occurs, the density exceeds the value 1.0 on the ordinate. It is possible to see from
238 Fig. 4 that the density of the different measurement configurations exceeds the unit
239 value in the range 2.0 to 2.5 dBA, meaning that the mode and the highest concentration
240 of values of differences in sound levels between the two microphones for the different
241 configurations are located in this range (Fig. 4).

242 a)



243

244 b)



245

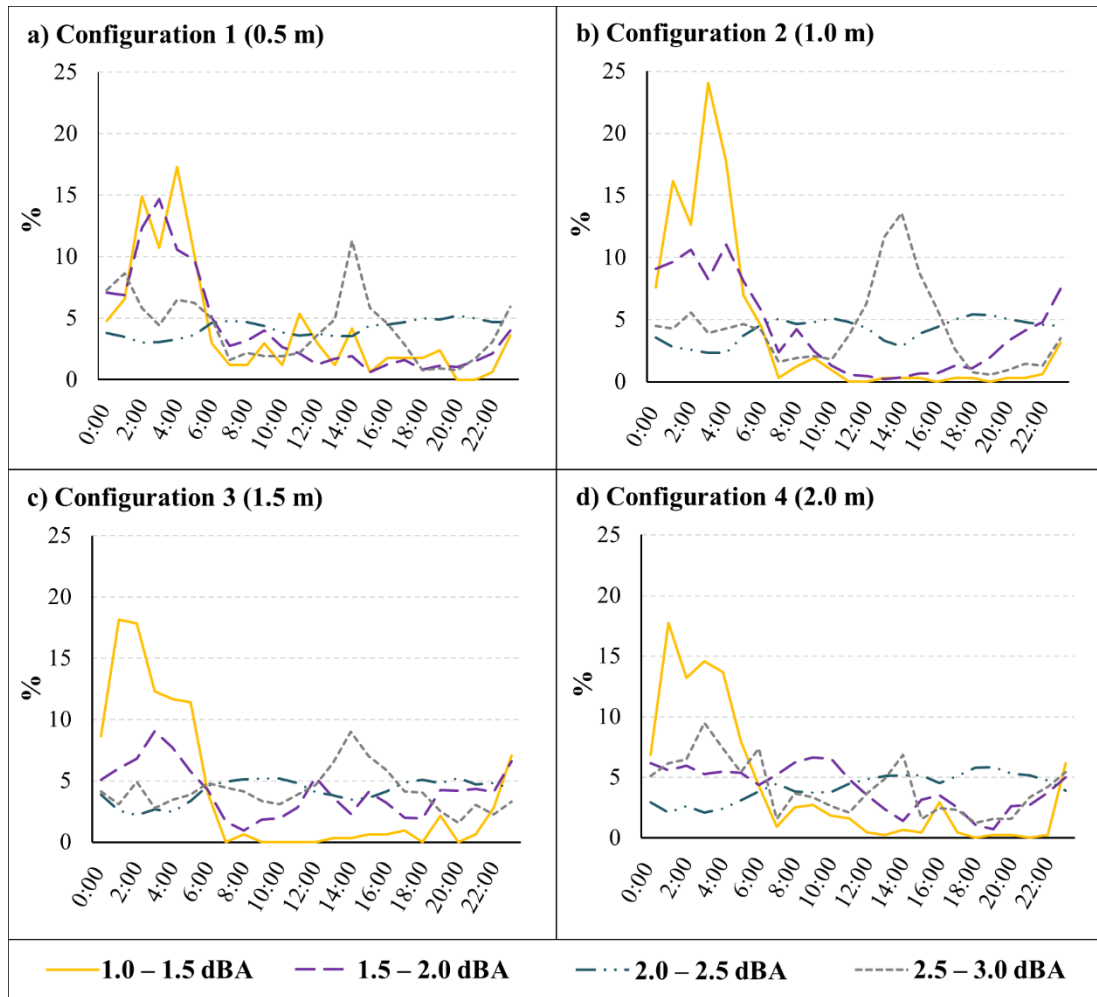
246 Fig. 4. Density of the differences in the sound levels registered by both
 247 microphones with an integration time of one minute over one week for each
 248 measurement configuration: a) 100% of the sound differences; b) sound differences
 249 between 1.0 and 3.0 dBA (95% of the data)

250 The distribution of the differences in sound levels for the different configurations
 251 seems similar (Fig. 4a). However, in a more detailed analysis in the range between 1.0
 252 and 3.0 dBA (Fig. 4b), where more than 95% of the data were concentrated, differences
 253 between the different configurations can be observed. As a result, variations in the
 254 position of the maximum in the distribution and asymmetries can be seen.

255 In order to improve the understanding of these differences in the density of the
 256 different configurations (Fig. 4b), the distributions of the sound differences were then
 257 analysed based on the time interval (Fig. 5) and the value of the registered sound level
 258 (Fig. 6).

259 The differences in the sound levels measured by the two microphones in four
260 intervals (1.0–1.5, 1.5–2.0, 2.0–2.5 and 2.5–3.0 dBA) are represented in Fig. 5 for each
261 measurement configuration. As mentioned above, these intervals contain more than
262 95% of the total data. The ordinate in Fig. 5 represents the proportion of data in a given
263 interval of differences considered that was measured in the time represented on the
264 abscissa.

265 It should be noted that the frequency distributions shown in Fig. 5 are visually
266 different depending on the range of differences considered (1.0–1.5, 1.5–2.0, 2.0–2.5
267 and 2.5–3.0 dBA) for all configurations. The range of time from 0:00 to 6:00 contains
268 the lowest values of the differences between 1.0 and 2.0 dBA. The percentage within
269 the 1.0 to 1.5 dB range is particularly notable. This result can be related to a previous
270 study [18] that suggested that road traffic would cease to be the main source of noise
271 during the night and that the sound level registered by microphones would correspond
272 to urban background noise, which would influence the difference in the levels registered
273 by both microphones. It should also be observed that the range 2.0 to 2.5 dBA is the
274 most stable and uniform throughout the day for all configurations. In addition, it is
275 interesting to note a small increase in the interval 2.5 to 3.0 dBA in the period of the day
276 between 12:00 and 17:00. This is centred at 14:00 and corresponds to the period of the
277 day with the highest traffic flow. This interval (2.5 to 3.0 dBA) also shows peaks at
278 night, mainly in the configurations 1 (Fig. 5a) and 4 (Fig. 5d). In the latter
279 configuration, the peak is higher than during the day. Finally, a comparison between the
280 different measurement configurations shows that configuration 4 (Fig. 5d) has the most
281 uniform distribution of differences throughout the day.



282

283 Fig. 5. Frequency distribution of the differences in sound levels registered minute by
 284 minute by both microphones with respect to the time interval of the day for each
 285 measurement configuration.

286 An analysis of the results in Fig. 5 suggests that the values of the measured sound
 287 levels may influence the differences found between both microphones. To analyse this
 288 possibility, Fig. 6 shows the proportion of data in the various ranges of sound level
 289 differences between both microphones (1.0–1.5, 1.5–2.0, 2.0–2.5 and 2.5–3.0 dBA),
 290 where the value of the sound level measured at the façade is represented on the ordinate.

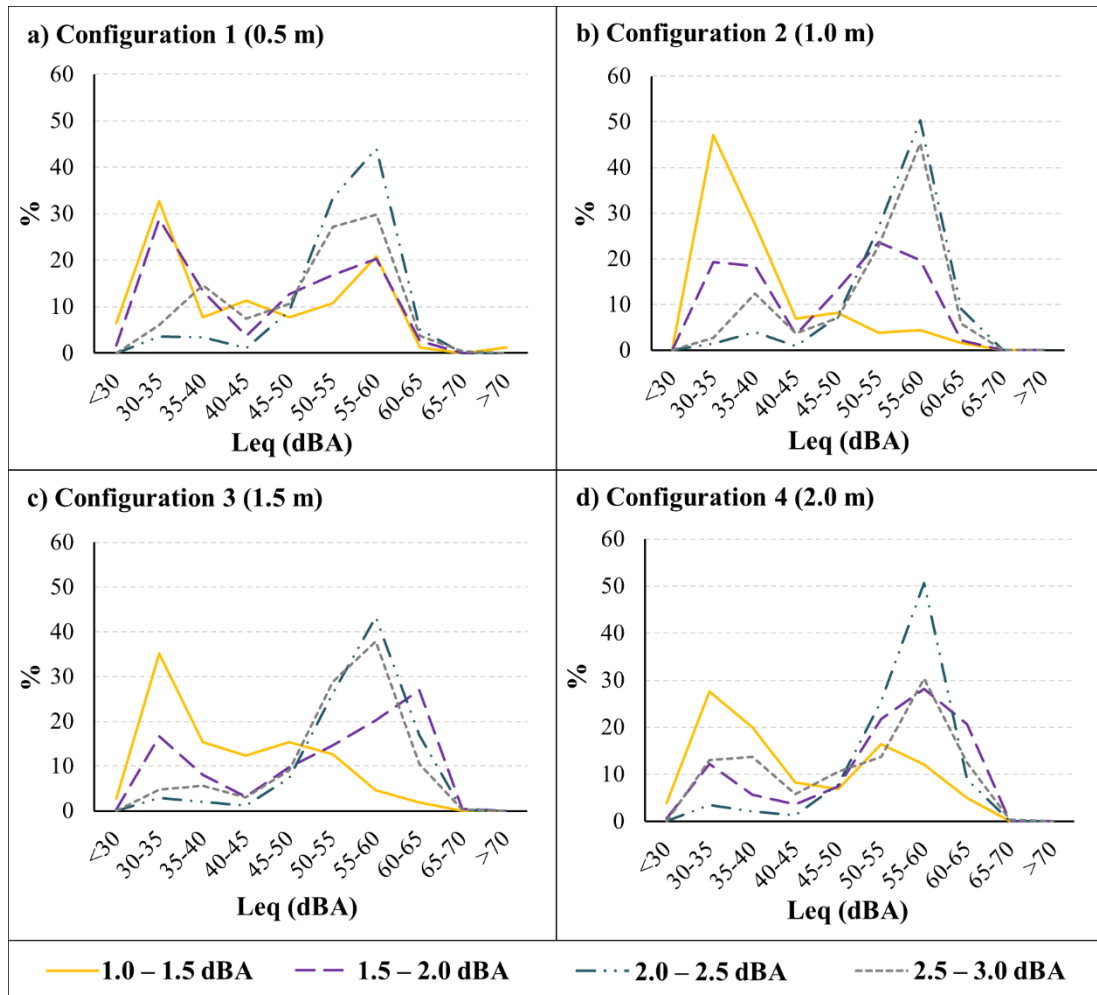
291 Fig. 6 shows that for each of the measurement configurations, each difference
 292 interval (1.0–1.5, 1.5–2.0, 2.0–2.5 and 2.5–3.0 dBA) has a different frequency
 293 distribution depending on the values of the measured sound levels.

294 The differences in the range 1.0–1.5 dBA show a structure in which the greatest
295 proportion of values fall into the range 30 to 45 dBA in all measurement configurations,
296 reaching maximum values in the range 30–35 dBA. Between 30% and 50% of the
297 measurements in which the difference between the two microphones is in the range 1.0–
298 1.5 dBA are concentrated in this interval. These results are in accordance with those
299 reported by Jagniatinskis et al. [18], and are related to the results shown in Fig. 5, in
300 which the highest proportions of the lowest values of the differences in levels are
301 registered from 0:00 to 6:00. In contrast, the range of differences 2.0–2.5 dBA shows a
302 very high proportion in the range of measured values from 50–65 dBA in all
303 measurement configurations. It also has a maximum in the range 55–60 dBA, with
304 proportions varying between 40% and 50% of the data in the range of differences 2.0–
305 2.5 dBA. Both intervals of differences (1.0–1.5 and 2.0–2.5 dBA) therefore show
306 similar behaviour in all measurement configurations.

307 However, the difference ranges of 1.5–2.0 dBA and 2.5–3.0 dBA present a more
308 variable frequency distribution than the previous ones depending on the sound level
309 measured in the façade and the measurement configuration used. In particular, the range
310 2.5–3.0 dBA shows a behaviour that is similar to that shown in the range 2.0–2.5 dBA,
311 but with significant percentages in the range of low measured values (30–45 dBA),
312 especially in configuration 4 (Fig. 6d). This result seems to confirm those reported
313 above, depending on the time of measurement; that is, low values of the sound levels
314 can also cause differences in the range 2.5–3.0 dBA. On the other hand, the difference
315 range 1.5–2.0 dBA contains high proportions for all configurations in the ranges 30–45
316 and 50–65 dBA.

317 If the results obtained for the different measurement configurations are compared,
318 it can be seen that the extreme values of the range of differences (1.0–1.5 dBA) are

319 typically found for low values of the measured sound levels in all configurations. In
 320 contrast, the highest values of the differences (in the range 2.5–3.0 dBA) are mainly
 321 found in the range of measured values of 50–65 dBA, although it is not negligible the
 322 proportion in the range of low sound levels (35–45 dBA) in configuration 4.



323

324 Fig. 6. Frequency distribution of the of the sound differences measured minute by
 325 minute over one week with respect to the sound levels measured at the façade for each
 326 measurement configuration.

327 Regarding the measurement with a microphone mounted on the façade, the broadband
 328 analysis revealed that measurements carried out with microphones at different distances
 329 from the reflecting surface differ from those indicated by the ISO 1996-2 standard. This

330 analysis also demonstrated that the distance at which the microphone is placed
 331 influences the result. A detailed temporal analysis shows that the lower values of the
 332 differences in sound levels measured by both microphones are concentrated at night in
 333 all configurations. It also can be concluded that it is particularly configuration 4 shows
 334 the greatest temporal stability of the differences. Furthermore, when the differences are
 335 related to the values of the measured sound levels, it was clearly detected that the lowest
 336 values of the differences are found for the lowest sound levels (30–45 dBA). On the
 337 other hand, when the differences are around the distribution modes (Fig. 4), the values
 338 of the sound levels are in the range 50–65 dBA.

339 **3.2. Spectral results**

340 The general frequency range proposed by the ISO 1996-2 standard includes the
 341 octave bands between 63 Hz and 8 kHz. However, for the case in which a microphone is
 342 placed directly on a surface, Annex B of ISO 1996-2 indicates that a 6 mm microphone
 343 should be used for octave-band measurements if the frequency range is expanded above
 344 4 kHz. Therefore, given that Brüel & Kjær 4950 microphones of the sound level meters
 345 used in this studio are 13 mm in diameter, a spectral study was carried out in octave
 346 bands between 63 Hz and 4 kHz (Fig. 7).

347 The spectrum of the differences between the two microphones used in each of the
 348 four configurations studied in the frequency range 63 Hz–4 kHz (Fig. 7) shows
 349 statistically significant differences between all pairs of configurations studied, as
 350 revealed by a multivariate analysis of variance (MANOVA) (Table 2).

351 Table 2. MANOVA for ΔL_{eq} (dB) in octave bands between pairs of configurations

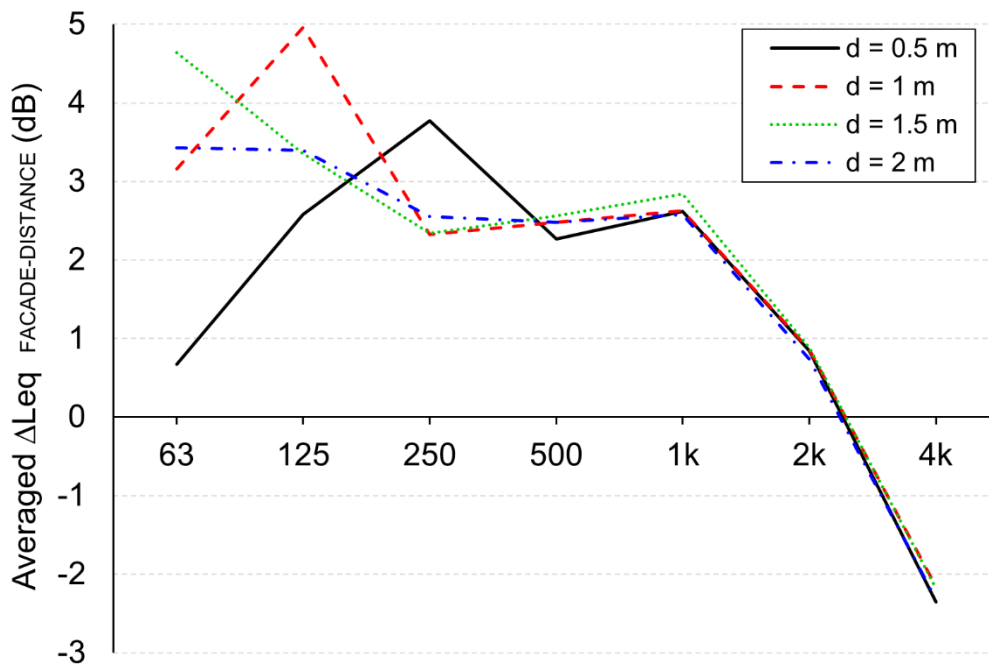
Configurations	Pillai's value	F	P-value
1–2	0.95	815.78	< 0.001

1-3	0.90	363.06	< 0.001
1-4	0.83	210.26	< 0.001
2-3	0.86	265.57	< 0.001
2-4	0.66	83.32	< 0.001
3-4	0.63	71.13	< 0.001

352

353 If octave bands are independently examined (Fig. 7), several results of interest can

354 be found.



355

356 Fig. 7. Average difference in equivalent sound level between both microphones in

357

frequency octave bands

358

359

360

361

362

363

When the high-frequency zone is evaluated, a very significant high negative in the differences between the two microphones can be observed at 4 kHz, regardless of the measurement configuration. In fact, this negative value indicates that the microphone located at a certain distance from the façade measures a higher sound level than that mounted at the façade. Since this result is the same for all four configurations, this minimum would be related to the existence of a destructive interference for a high

364 frequency for the microphone located at the reflecting surface. In addition, it can also be
 365 noted that the value of the differences between the two microphones at a frequency band
 366 of 2 kHz is different in all configurations from the expected value of 2.7 dB based on
 367 the ISO 1996-2 standard. Therefore, both for the 4 kHz and 2 kHz bands, the location of
 368 a microphone at the façade would cause the incident sound levels for buildings to be
 369 assessed at a lower level than the real value, meaning that the doses of noise received at
 370 the façades at these frequencies would be underestimated. Variations in these
 371 frequencies influence the broadband levels in dBA that are used to determine the effects
 372 of sound exposure on the health of the resident population.

373 Table 3. *P*-values for pairwise comparisons of configurations using a *t*-test for ΔL_{eq}
 374 (dB) in octave bands

ΔL_{eq} (dB)	Configuration	1	2	3
63 Hz	2	< 0.001	-	-
	3	< 0.001	< 0.001	-
	4	< 0.001	> 0.05	< 0.001
125 Hz	2	< 0.001	-	-
	3	< 0.001	< 0.001	-
	4	< 0.001	< 0.001	> 0.05
250 Hz	2	< 0.001	-	-
	3	< 0.001	> 0.05	-
	4	< 0.001	< 0.001	< 0.001
500 Hz	2	< 0.001	-	-
	3	< 0.001	< 0.001	-
	4	< 0.001	> 0.05	< 0.001
1 kHz	2	> 0.05	-	-
	3	< 0.001	< 0.001	-
	4	> 0.05	> 0.05	< 0.001
2 kHz	2	> 0.05	-	-
	3	> 0.05	> 0.05	-

	4	> 0.05	< 0.001	< 0.001
	2	> 0.05	-	-
4 kHz	3	> 0.05	> 0.05	-
	4	> 0.05	> 0.05	> 0.05

375

376 Turning to the low- and medium-frequency range, from an analysis of the results
377 shown in Fig. 7 it can be concluded that both destructive and constructive interference
378 phenomena can occur at the microphone located at a certain distance from the façade
379 depending on the selected distance. A destructive interference phenomenon is found for
380 configurations 1, 2 and 3 at frequencies that depend on the distance between the
381 microphone and the reflecting surface. At a distance of 0.5 m (configuration 1), a
382 maximum value of the difference in the sound levels is obtained at 250 Hz. This
383 maximum occurs at 125 Hz for configuration 2 and 63 Hz for configuration 3. For
384 configuration 4, in which the microphone is located 2.0 m from the façade, this
385 maximum does not appear. This result is related to that specified by the ISO 1996-2
386 standard in Annex B (informative) with regard to the appearance of coherence effects at
387 low frequencies and the recommendation of a minimum distance of 1.6 m for
388 measurements in octave bands. However, it is shown in Fig. 7 that for an octave band of
389 63 Hz at a distance of 1.5 m, detectable interference remains between the direct and
390 reflected sound waves. On the other hand, a decrease in the differences can be seen at
391 low frequencies in terms of the appearance of constructive interference. This is detected
392 at 63 Hz in the case of configuration 1, and must be related to the existence of
393 constructive interference at 0.5 m between the direct and reflected waves from the
394 façade. These results are also linked to the interference effects of waves reported in
395 research by Quirt et al. [17] and Hopkins et al. [48].

396 An analysis of the results for the frequency bands in Table 3 concerning the
 397 significance of the differences shows that: a) the low-frequency bands from 63 Hz to
 398 500 Hz generally show significant differences in most of the comparisons we can make
 399 between pairs of configurations; b) in the medium-frequency bands between 500 Hz and
 400 1000 Hz, the 500 Hz band shows a behaviour in terms of the significance of the
 401 differences that is similar to that at low frequencies when configurations are compared;
 402 and the 1 kHz band behaves in an intermediate way between the low- and high-
 403 frequency bands; c) in general, the high-frequency bands of 2 and 4 kHz do not show
 404 significant differences for the different configurations.

405 Based on the results of the above analysis of the spectrum of differences, prior
 406 works concerning the suitability of the use of A-weighting to assess the annoyance of
 407 environmental noise [29,31] and the possibility that A-weighting is a poor predictor of
 408 annoyance due to low-frequency noise [30,33], it is of interest to assess the influence of
 409 the range frequencies considered in the measurements, as indicated in ISO 1996-2
 410 standard, and the use of weighting in the evaluation of continuous equivalent sound
 411 levels. Table 4 below shows the average difference in equivalent broadband sound
 412 levels for the different octave frequency band ranges, in dB, dBA and dBC.

413 If the same analysis is carried out by configuration, it can be observed that
 414 configuration 1 differs from the others in all low-frequency bands up to 500 Hz.
 415 Configuration 2 differs from configuration 3, and configuration 3 from configuration 4
 416 generally at low and medium frequencies.

417 Table 4: Differences in equivalent broadband levels obtained at different frequency
 418 ranges and with different frequency weightings (Z, A and C)

Config.	Frequency ranges for ΔL_{eq}		
	63–8000 Hz	63–4000 Hz	125–4000 Hz

	Z (dB)	A (dBA)	C (dBC)	Z (dB)	A (dBA)	C (dBC)	Z (dB)	A (dBA)	C (dBC)
1	1.5	2.2	1.5	1.5	2.2	1.5	2.6	2.2	2.6
2	3.1	2.3	3.1	3.1	2.3	3.1	3.0	2.3	3.0
3	3.6	2.3	3.5	3.6	2.3	3.5	2.7	2.3	2.7
4	3.0	2.2	3.0	3.0	2.2	3.0	2.6	2.2	2.6

419

420 It is worth noting that for A-weighting, there are no variations between the
421 differences of the measured sound levels in both microphones for any of the
422 configurations analysed when the 8 kHz or 63 Hz bands are eliminated (Table 4). In
423 addition, a strong similarity between the results obtained using linear or C-weighting
424 can be noted.

425 If now the linear and C configurations are analysed, Table 4 shows that removing
426 the 8 kHz band has no detectable effect on the global level in any configuration when
427 these weightings are used. This is undoubtedly associated with the low sound energy of
428 the 8 kHz band in the measured noise spectrum. However, if the 63 Hz band is
429 eliminated, an important effect arises, indicating that the sound energy in that frequency
430 band is important for the measured noise. In configuration 1, the variation between the
431 global values is greater than 1 dB, and this is probably associated with the existence of a
432 constructive interference in the 63 Hz band for the microphone located 0.5 m from the
433 façade. This variation is only 0.1 dB for configuration 2, probably due to the great
434 weight of destructive interference for the microphone located at 1.0 m at a frequency of
435 125 Hz. For configuration 3, the difference is a decrease of slightly less than 1 dB, and
436 this must be associated with destructive interference at that frequency in the microphone
437 located at 1.5 m from the façade. Finally, a moderate decrease in the difference of 0.4
438 dB between microphones is still detected for configuration 4, due to the weight of this
439 band in the global differences.

440 The variations in the low- and medium-frequency range shown in Fig. 7 are
441 therefore transferred to the overall results using linear weighting. This effect is stronger
442 in configurations 1 and 3. However, all the differences detected according to the
443 considered spectral range do not appear when A-weighting is applied. Consequently, the
444 attenuation due to the use of A-weighting in the weights of low frequencies eliminates
445 the differences in the spectrum that arise when measurements are carried out at different
446 distances from the façade. Therefore, if the evaluation of the noise doses received at low
447 frequency is of interest and the use of C-weighting is adequate, the location of the
448 microphone with respect to the façade may be of great relevance based on the effects
449 that arise [32,33] or if the possible problems analysed in the introduction due to the use
450 of A-weighting [10, 29-31] are taken into account.

451 Despite the important differences between the results obtained in this analysis of
452 the noise spectrum for the octave bands in the range 63 Hz to 500 Hz, as a conclusion of
453 this study it can be pointed out that for the measurement of traffic noise, given its
454 spectrum and the effects of A-weighting, if somebody wishes to assess the overall noise
455 level incident on the façade, the range 0.5 to 2 m indicated by the ISO 1996-2 standard
456 for location of the microphone leads to similar results. In the present study, the
457 differences between the façade and mobile microphones are found to be always in the
458 low range of the error interval indicated by the standard.

459 If the spectral characteristics of the measured noise are different from traffic
460 noise, the distance between the façade and the location of the microphone may become
461 relevant. In this case, measurement at 2 m would be the best option, especially if low
462 frequency noise is expected. In addition, in order to accurately assess the sound level
463 incident on the façade, both in terms of the overall values and the spectrum, the effects
464 detected in the measurements with a microphone mounted on the façade suggest that the

465 most appropriate option is to place the microphone at 2 m from the façade, as well as
466 meeting the indications given by the ISO 1996-2 standard.

467 **4. CONCLUSIONS**

468 An analysis was carried out in this study of the differences between the sound
469 values measured in broadband and octave bands by a microphone placed at a certain
470 distance from the façade (0.5, 1.0, 1.5 and 2.0 m) and another placed at the façade.
471 Measurements were made over one week for each configuration following the
472 guidelines of ISO 1996-2 standard for the microphone locations relative to reflecting
473 surfaces (Annex B).

474 The following results were obtained from the broadband analysis of the
475 differences in the measured sound levels:

- 476 a) Under the conditions of this study, average differences in sound levels
477 experimentally determined were lower than the expected value of 2.7 dBA
478 indicated by the ISO 1996-2 standard. These differences were found at the lower
479 limit of uncertainty, or even outside the range indicated by the ISO 1996-2
480 standard. Considering the standard deviation, these average differences did not
481 include the value of 2.7 dB.
- 482 b) An analysis of these differences and the sound levels recorded with the façade
483 microphone showed that the smallest differences were found when the sound
484 levels were lowest (30–45 dBA). However, the differences were around the
485 average values when the measured sound levels were in the range 50–65 dBA.
- 486 c) It was detected in this study that the smallest differences (1.0-1.5 dB) between
487 the sound levels measured by the microphone mounted on the façade and the
488 microphone located at different distances are concentrated in the night time
489 between 1:00 and 6:00, when the flow of vehicles decreases.

490 From the analysis of the differences in octave bands, the following conclusions
491 can be drawn:

492 a) Significant differences were found between the configurations in terms of octave
493 bands. Variations greater than 2 dB were obtained in the bands of 63, 125 and
494 250 Hz, which were associated with constructive or destructive interference
495 phenomena depending on the distance from the microphone to the façade. It is
496 therefore important that the measurement configuration is taken into account in
497 the assessment of the doses of noise in low frequencies.

498 b) Destructive interference phenomena were detected at the façade microphone in
499 the octave bands of 2 and 4 kHz. This indicates that the dose of noise received
500 would be underestimated at these frequencies for this measurement location.

501 c) For traffic noise, the use of A-weighting eliminates the effects of the differences
502 in the spectrum when the global results are analysed. However, if a C-weighting
503 is used, the variations in the spectrum are transferred to the broadband noise.

504 Finally, considering the results of this work and the distance range established in
505 ISO 1996-2 for the position of the microphone in front of the reflecting surface, a
506 microphone location of 2 m from the façade can be considered the best option for an
507 accurate evaluation of the sound level incident on the façade if the indications of the
508 ISO 1996-2 standard can be met.

509

510 **Acknowledgements**

511 The authors wish to thank the funded projects TRA2015-70487-R
512 (MINECO/FEDER, UE) and the European Regional Development Fund (ERDF). This
513 research has been supported by project GR10175 (Junta de Extremadura,
514 Spain/European Regional Development Funds, EU) and by the Chilean National

515 Commission for Scientific and Technological Research (CONICYT) through the
516 project FONDECYT No. 1180547. David Montes González was supported by
517 Consejería de Economía, Ciencia y Agenda Digital of Junta de Extremadura, European
518 Union and European Social Fund (ESF), through grants for the strengthening of
519 R&D&I by means of the mobility of postdoctoral researchers (PO17014). Guillermo
520 Rey Gozalo was supported by Juan de la Cierva–Incorporación contract from the
521 Spanish Ministry of Economy, Industry and Competitiveness (IJCI-2016-28923), and
522 Consejería de Economía, Ciencia y Agenda Digital of Junta de Extremadura, through
523 grants for attracting and returning research talent to R&D&I centres belonging to the
524 Extremadura Science, Technology and Innovation System (TA18019).

525

526 REFERENCES

- 527 [1] END. Directive 2002/49/EC of the European Parliament and of the Council of 25
528 June 2002 relating to the assessment and management of environmental noise
529 (END). Brussels: The European Parliament and the Council of the European
530 Union; 2002.
- 531 [2] WG-AEN. Good Practice Guide for Strategic Noise Mapping and the Production
532 of Associated Data on Noise Exposure. European Commission Working Group-
533 Assessment of Exposure to Noise (WG-AEN); version 2, 13th August 2007.
- 534 [3] Paschalidou AK, Kassomenos P, Chonianiaki F. Strategic Noise Maps and Action
535 Plans for the reduction of population exposure in a Mediterranean port city. *Sci*
536 *Total Environ.* 2019;654:144–53. doi:[https://doi.org/10.1016/j.scitotenv.2018.11.](https://doi.org/10.1016/j.scitotenv.2018.11.048)
537 048
- 538 [4] Sánchez-Sánchez R, Fortes-Garrido JC, Bolívar JP. Noise monitoring networks as
539 tools for smart city decision-making. *Arch Acoust* 2018;43(1):103–112.

- 540 doi:10.24425/118085
- 541 [5] Dintrans A, Préndez M. A method of assessing measures to reduce road traffic
542 noise: A case study in Santiago, Chile. *Appl Acoust* 2013;74:1486–91.
543 doi:<https://doi.org/10.1016/j.apacoust.2013.06.012>
- 544 [6] Commission Directive (EU) 2015/996 of 19 May 2015 establishing common
545 noise assessment methods according to Directive 2002/49/EC of the European
546 Parliament and of the Council. Brussels: The European Parliament and the
547 Council of the European Union; 2015.
- 548 [7] ISO 1996-1. Description, measurement and assessment of environmental noise.
549 Part 1: Basis quantities and assessment procedures. Geneva: International
550 Organization for Standardization; 2016.
- 551 [8] ISO 1996-2: 2017. Description, measurement and assessment of environmental
552 noise. Part 2: Determination of sound pressure levels. Switzerland: International
553 Organization for Standardization; 2017.
- 554 [9] Bartels S, Rooney D, Müller U. Assessing aircraft noise-induced annoyance
555 around a major German airport and its predictors via telephone survey – The
556 COSMA study. *Transp Res Part D: Transp and Environ* 2018;59:246–58.
557 doi:<https://doi.org/10.1016/j.trd.2018.01.015>
- 558 [10] Rey-Gozalo G, Barrigón-Morillas JM, Montes-González D, Atanasio-Moraga P.
559 Relationships among satisfaction, noise perception, and use of urban green spaces.
560 *Sci Total Environ* 2018;624:438–50. doi:[https://doi.org/10.1016/j.scitotenv.
561 2017.12.148](https://doi.org/10.1016/j.scitotenv.2017.12.148)
- 562 [11] Maijala P, Shuyang Z, Heittola T, Virtanen T. Environmental noise monitoring
563 using source classification in sensors. *Appl Acoust* 2018;129:258–67.
564 doi:<https://doi.org/10.1016/j.apacoust.2017.08.006>

- 565 [12] Montes-González D, Barrigón-Morillas JM, Godinho L, Amado-Mendes P.
566 Acoustic screening effect on building façades due to parking lines in urban
567 environments. *Effects in noise mapping. Appl Acoust* 2018;130:1–14.
568 doi:<https://doi.org/10.1016/j.apacoust.2017.08.023>
- 569 [13] Minichilli F, Gorini F, Ascari E, Bianchi F, Coi A, Fredianelli L, et al. Annoyance
570 Judgment and Measurements of Environmental Noise: A Focus on Italian
571 Secondary Schools. *Intern J Environ Res and Pub Health* 2018;15.
572 doi:10.3390/ijerph15020208
- 573 [14] Montes-González D, Barrigón-Morillas JM, Godinho L, Amado-Mendes P, Rey-
574 Gozalo G, Atanasio-Moraga P. Selection of microphone location, measurement
575 uncertainty and calculated noise maps. 24th International Congress on Sound and
576 Vibration; 2017.
- 577 [15] ISO 1996-2: 2007. Description, measurement and assessment of environmental
578 noise. Part 2: Determination of environmental noise levels. Switzerland:
579 International Organization for Standardization; 2007.
- 580 [16] Hall FL, Papakyriakou MJ, Quirt JD. Comparison of outdoor microphone
581 locations for measuring sound insulation of building façades. *J Sound Vib*
582 1984;92:559–67. doi:10.1016/0022-460X(84)90198-6
- 583 [17] Quirt JD. Sound fields near exterior building surfaces. *J Acoust Soc Am*
584 1985;77:557–66. doi:<https://doi.org/10.1121/1.391873>
- 585 [18] Jagniatinskis A, Fiks B. Assessment of environmental noise from long-term
586 window microphone measurements. *Appl Acoust* 2014;76:377–85.
587 doi:10.1016/j.apacoust.2013.09.007
- 588 [19] Memoli G, Paviotti M, Kephelopoulos S, Licitra G. Testing the acoustical
589 corrections for reflections on a façade. *Appl Acoust* 2008;69:479–95.

- 590 doi:10.1016/j.apacoust.2007.05.006
- 591 [20] Montes González D, Barrigón Morillas JM, Rey Gozalo G. The influence of
592 microphone location on the results of urban noise measurements. *Appl Acoust*
593 2015;90:64–73. doi:10.1016/j.apacoust.2014.11.001
- 594 [21] Mateus M, Dias Carrilho J, Gameiro Da Silva M. An experimental analysis of the
595 correction factors adopted on environmental noise measurements performed with
596 window-mounted microphones. *Appl Acoust* 2015;87:212–8.
597 doi:10.1016/j.apacoust.2014.06.019
- 598 [22] Flores R, Asensio C, Gagliardi P, Licitra G. Study of the correction factors for
599 aircraft noise façade measurements. *Appl Acoust* 2019;145:399–407.
600 doi:10.1016/j.apacoust.2018.10.007
- 601 [23] Barrigón Morillas JM, Montes González D, Rey Gozalo G. A review of the
602 measurement procedure of the ISO 1996 standard. Relationship with the European
603 Noise Directive. *Sci Total Environ* 2016;565:595–606. doi:10.1016/j.scitotenv.
604 2016.04.207
- 605 [24] Montes Gonzalez D, Barrigón Morillas JM, Rey Gozalo G. Acoustic behaviour of
606 plates made of different materials for measurements with the microphone flush
607 mounted. *Appl Acoust* 2018;132:135–41. doi:10.1016/j.apacoust. 2017.11.011
- 608 [25] Mioduszewski P, Ejsmont JA, Grabowski J, Karpiński D. Noise map validation
609 by continuous noise monitoring. *Appl Acoust* 2011;72:582–9.
610 doi:https://doi.org/10.1016/j.apacoust.2011.01.012
- 611 [26] Szczodrak M, Kotus J, Kostek B, Czyżewski A. Creating Dynamic Maps of Noise
612 Threat Using PL-Grid Infrastructure. *Arch Acoust* 2013;38(2):235-242. doi:
613 https://doi.org/10.2478/aoa-2013-0028
- 614 [27] ISO 11819-4. Acoustics – Method for measuring the influence of road surfaces on

615 traffic noise – Part 4: SPB method using backing board. International
616 Organization for Standardization; 2013.

617 [28] ISO 10847. Acoustics - *In-situ* determination of insertion loss of outdoor noise
618 barriers of all types. International Organization for Standardization; 1997.

619 [29] Schomer PD, Suzuki Y, Saito F. Evaluation of loudness-level weightings for
620 assessing the annoyance of environmental noise. *J Acoust Soc Am* 2001;110(5
621 I):2390-2397. doi:<https://doi.org/10.1121/1.1402116>

622 [30] Nilsson ME. A-weighted sound pressure level as an indicator of short-term
623 loudness or annoyance of road-traffic sound. *J Sound Vib* 2007;302(1-2):197-207.
624 doi:<https://doi.org/10.1016/j.jsv.2006.11.010>

625 [31] Nilsson ME, Andéhn M, Lesna P. Evaluating roadside noise barriers using an
626 annoyance-reduction criterion. *J Acoust Soc Am* 2009;124(6):3561-3567.
627 doi:<https://doi.org/10.1121/1.2997433>

628 [32] Berglund B, Hassmén P, Soames Job RF. Sources and effects of low-frequency
629 noise. *J Acoust Soc Am* 1996;99(5):2985-3002.
630 doi:<https://doi.org/10.1121/1.414863>

631 [33] Persson Wayne K, Rylander R. The prevalence of annoyance and effects after long-
632 term exposure to low-frequency noise. *J Sound Vib* 2001;240(3):483-497.
633 doi:<https://doi.org/10.1006/jsvi.2000.3251>

634 [34] Barrigón Morillas JM, Gómez Escobar V, Méndez Sierra JA, Vílchez Gómez R,
635 Trujillo Carmona J. An environmental noise study in the city of Cáceres, Spain.
636 *Appl Acoust* 2002;63(10):1061-1070. doi:[https://doi.org/10.1016/S0003-
637 682X\(02\)00030-0](https://doi.org/10.1016/S0003-682X(02)00030-0)

638 [35] Barrigón Morillas JM, Escobar VG, Carmona JT, Sierra JAM, Vílchez-Gómez R,
639 Río FJCD. Analysis of the prediction capacity of a categorization method for urban

640 noise assessment. *Appl Acoust* 2011;72(10):760-771.
641 doi:<https://doi.org/10.1016/j.apacoust.2011.04.008>

642 [36] Rey-Gozalo, G.; Barrigón Morillas, J.M.; Gómez-Escobar, V. Analysing nocturnal
643 noise stratification. *Sci Total Environ* 2014;479-480 (1):39-47.
644 doi:10.1016/j.scitotenv.2014.01.130

645 [37] ANSI S12.18. Procedures for outdoor measurement of sound pressure level. New
646 York: Acoustical Society of America; 1994.

647 [38] Geraghty D, O'Mahony M. Investigating the temporal variability of noise in an
648 urban environment. *Int J Sustain Built Environ* 2016;5:34-45.
649 doi:<https://doi.org/10.1016/j.ijbsbe.2016.01.002>

650 [39] Juan Miguel Barrigón Morillas; Carmen Ortiz-Caraballo; Carlos Prieto Gajardo.
651 The temporal structure of pollution levels in developed cities. *Sci Total Environ*
652 2015; 517: 31-37. doi:10.1016/j.scitotenv.2015.02.057

653 [40] Quintero G, Balastegui A, Romeu J. Annual traffic noise levels estimation based
654 on temporal stratification. *J Environ Manag* 2018;206:1-9.
655 doi:10.1016/j.jenvman.2017.10.008

656 [41] Prieto Gajardo, C., Barrigón Morillas, J.M., Rey-Gozalo, G., Vilchez-Gómez, R.
657 Can weekly noise levels of urban road traffic, as predominant noise source,
658 estimate annual ones? *J Acoust Soc Am* 2016;140 (5): 3702-3709.
659 doi:<http://dx.doi.org/10.1121/1.4966678>

660 [42] Prieto Gajardo, C; Barrigón Morillas, J.M. Stabilisation patterns of hourly urban
661 sound levels. *Environ Mon Assess* 2015;187(1): 4072-4987. doi:10.1007/s10661-
662 014-4072-3

663 [43] Can, A., Van Renterghem, T., Rademaker, M., Dauwe, S., Thomas, P., De Baets,
664 B. Sampling approaches to predict urban street noise levels using fixed and

665 temporary microphones. *J Environ Monit* 2011;13(10):2710–2719.
666 doi:10.1039/c1em10292c

667 [44] Zambon, G., Benocci, R., Bisceglie, A., Roman, H.E. Milan dynamic noise
668 mapping from few monitoring stations: Statistical analysis on road network.
669 Proceedings of 45th INTER-NOISE 2016; 6350-6361

670 [45] Smiraglia, M., Benocci, R., Zambon, G., Roman, H.E. Predicting hourly traffic
671 noise from traffic flow rate model: Underlying concepts for the DYNAMAP
672 project. *Noise Mapping* 2016; 3(1);130-139. doi: 10.1515/noise-2016-0010

673 [46] Zambon, G., Roman, H.E., Smiraglia, M., Benocci, R. Monitoring and prediction
674 of traffic noise in large urban areas. *Appl Sci.* 2018;8(2);251-267.
675 doi:<https://doi.org/10.3390/app8020251>

676 [47] Wei W, Renterghem TV, Coensel BD, Botteldooren D. Dynamic noise mapping: A
677 map-based interpolation between noise measurements with high temporal
678 resolution. *Appl Acoust.* 2016;101:127–40.
679 doi:<https://doi.org/10.1016/j.apacoust.2015.08.005>

680 [48] Hopkins C, Lam Y. Sound fields near building façades- comparison of finite and
681 semi- infinite reflectors on a rigid ground plane. *Appl Acoust* 2009;70(2); 300-308.
682 doi:<https://doi.org/10.1016/j.apacoust.2008.03.008>