

2

3 **Short and long-term effects of different irrigation and tillage systems on soil properties and**

4 **rice productivity under Mediterranean conditions**

5 Javier Sánchez-Llerena^{a*} Antonio López-Piñeiro^a, Ángel Albarrán^b, David Peña^a, Daniel

6 Becerra^b, José Manuel Rato-Nunes^c.

7 ^a *Department of Plant Biology, Ecology and Earth Sciences, Faculty of Sciences, University of*

8 *Extremadura, Avda de Elvas s/n, 06071 Badajoz, Spain.*

9 ^b *Department of Forestry and Agricultural Engineering, University of Extremadura, Avda Adolfo*

10 *Suárez s/n, 06071 Badajoz, Spain.*

11 ^c *Elvas Agricultural School, Polytechnic Institute of Portalegre, Avda 14 de Janeiro s/n, 7350*

12 *Elvas, Portugal.*

13 **Abstract**

14 In Mediterranean environments, flood irrigation of rice crops is in danger of
15 disappearance due to its unsustainable nature. The aim of the present study was to determine
16 the short- and long-term effects of aerobic rice production, combined with conventional and
17 no-tillage practices, on soils' physical, physicochemical, and biological properties, as well as on
18 the rice yield components and productivity in the semi-arid Mediterranean conditions of SW
19 Spain. A field experiment was conducted for three consecutive years (2011, 2012, and 2013),
20 with four treatments: anaerobic with conventional tillage and flooding (CTF), aerobic with
21 conventional tillage and sprinkler irrigation (CTS), aerobic with no-tillage and sprinkler
22 irrigation (NTS), and long-term aerobic with no-tillage and sprinkler irrigation (NTS7).
23 Significant soil properties improvements were achieved after the long-term implementation of
24 no-tillage and sprinkler irrigation (NTS7). The short-term no-tillage and sprinkler irrigated

25 treatment (NTS in 2011 and 2012) gave lower yields than CTF, but reached similar yields in the
26 third year (NTS 8229 kg ha⁻¹; CTF 8926 kg ha⁻¹), with average savings of 75% of the total
27 amount of water applied in CTF. The NTS7 data showed that high yields (reaching 9805 kg ha⁻¹
28 in 2012) and water savings are sustainable in the long term. The highest water productivity
29 was with NTS7 in 2011 (0.66 g L⁻¹) and 2012 (1.46 g L⁻¹), and with NTS in 2013 (1.05 g L⁻¹). Thus,
30 mid- and long-term implementation of sprinkler irrigation combined with no-tillage may be
31 considered as a potentially productive and sustainable rice cropping system under
32 Mediterranean conditions.

33 **Keywords:** aerobic rice, sprinkler irrigation, no-tillage, water productivity.

34

35 1. Introduction

36 Rice is a staple food for more than 50% of the world's population, reaching up to 80%
37 when considering only the Asian population. Rice production must increase by 70% to meet
38 demand by 2050 as the world population keeps increasing (Wu *et al.*, 2013). This crop takes up
39 to 11% of world farmland (180 million ha). This figure is 88% in the case of Asia which accounts
40 for 90% of global rice production (750 million tons of paddy rice). Besides its importance as a
41 source of food, rice growing is the largest employment sector of rural populations in Asia, and
42 the crop is also widely cultivated in Africa and America, and intensively in some parts of
43 southern Europe, mainly in Mediterranean environments (CGIAR, 2006). In the European
44 Union (EU), 475 000 ha of land are devoted to rice growing, with a production of 3.2 millions of
45 tons of paddy rice. Spain is one of the largest rice producers in the EU, accounting for 20% of
46 total European rice crop farmland, and 30% of total rice production. Within Spain,
47 Extremadura, with an average yield of 7300 kg ha⁻¹, has become consolidated as one of the
48 largest rice producing regions, with the most productive areas concentrated in the River
49 Guadiana's historical floodplains (MAGRAMA, 2013).

50 World rice production has suffered for years from an important lack of investment
51 regarding research on and development of new management techniques, slowing down the
52 implementation of beneficial innovations (Van Tran, 2002). Environmentally, modern
53 approaches to rice crop intensification have damaged important natural resources, causing
54 significant increases in soil salinity, water pollution, health problems from high pesticide
55 concentrations in water and food, and increased greenhouse gas emissions (Devkota *et al.*,
56 2015). The agriculture sector is highly exposed to the risks accompanying climate change
57 (Lanfranchi *et al.*, 2014). Therefore, possible global climate changes associated with increased
58 atmospheric concentrations of greenhouse gases are likely to affect the efficiency of
59 agricultural production systems (Porter, 2005). Together with this, one of the major challenges

60 facing rice farming is to produce the same or more with less water, labour, and chemicals, so
61 as to ensure the sector's long-term sustainability (Khumar & Ladha, 2011). In order to avoid
62 long-term supply and demand imbalances, innovations are needed in agronomic management
63 and technology. In particular, the development of alternative rice farming systems may
64 contribute to reduce the aforementioned negative effects and to increase productivity
65 (Mukhopadhyay *et al.*, 2013). Research on these management techniques needs to mainly
66 focus on a rational and sustainable use of two essential resources: soil and water.

67 Rice crop cultivation in Mediterranean environments traditionally involves intensive and
68 expensive tillage practices. The leveling, ploughing, cross-harrowing, and final puddling may be
69 direct causes of dramatic decreases in soil quality, initiating processes that may change for the
70 worse the soil's original physical properties. These result in organic matter and nutrient
71 depletion, increased penetration resistance due to the creation of plough pan and surface
72 crust, as well as increased acidity and reduced microbial activity (Bezdicsek *et al.*, 2003).
73 Furthermore, new studies carried out under Mediterranean conditions (Moreno-Jiménez *et al.*,
74 2014) show an important increase in heavy metal content in both soil and grain after several
75 years of rice monoculture using conventional tillage practices and flood irrigation.

76 There are currently many proposals regarding soil conservation in intensive agriculture
77 environments. Most of them are centred on the concept of "conservation agriculture".
78 Conservation agriculture is a modern alternative for resource-saving crop production that
79 strives to achieve acceptable profits from high and sustained production levels while
80 concurrently conserving the environment. It is characterized by three interlinked principles:
81 minimal mechanical soil disturbance, permanent organic soil cover, and diversification of
82 either crop rotations in the case of annual crops or plant associations in the case of perennial
83 crops. These principles reach their greatest expression with the technique known as no-tillage,

84 extensively used to increase soil organic matter, to control soil degradation, and to increase
85 water holding capacity (López-Garrido *et al.*, 2014).

86 Water consumption in traditional rice production systems in Spain is on average
87 24 000 m³ ha⁻¹yr⁻¹. In a semi-arid Mediterranean environment, where water resources are
88 strongly limited, flood irrigation of rice crops is in danger of disappearance due to its
89 unsustainable nature. The continual growth of the population and the consequent growing
90 need for drinking water is a global problem, so that any way to save water is of great
91 importance (Yan *et al.*, 2015). In addition, available water resources for human consumption
92 have been reduced by the competing demands of industry, so that there is a need for the
93 development of new alternatives that allow more efficient water use in agriculture. Rice
94 farming is obviously one of the main objectives to reduce water consumption since it accounts
95 for some 50% of irrigation water used worldwide (Barker *et al.*, 1999). In this context, water
96 productivity has taken on an importance that is even greater than that of yield itself, so that
97 reducing water consumption and increasing its productivity constitute one of the main
98 challenges that the rice sector must face in the near future. Since, however, rice is very
99 sensitive to water stress, a reduction in the total amount of irrigation water used may lead to a
100 major decrease in yield, threatening the economic viability of the crop at its present levels.

101 Aerobic (non-flooded) rice is one of the new water-saving production systems being
102 developed all over the world. In this, rice is grown on well-drained, unsaturated soils
103 (Rajakumar *et al.*, 2009). Hence, irrigation needs are reduced because of reduced water losses
104 to percolation, evaporation, and surface run-off. This system may be considered an especially
105 interesting alternative for semi-arid Mediterranean environments (Facchi *et al.*, 2013).
106 However, the development of this system is still in its early stages, and more research is
107 needed in order to correctly implement these techniques. Thus, sprinkler irrigation has been
108 proposed as an efficient management technique for aerobic rice systems in semi-arid

109 Mediterranean conditions. In Europe, research on sprinkler irrigated rice has been limited to
110 the typical paddy fields of northern Italy (Russo & Nardi, 1996) and some irrigated areas of
111 southern Italy and Sardinia (Spanu *et al.*, 1997), giving 50% of irrigation water savings
112 compared with traditional flood irrigation. Recent studies have given promising results when
113 applying no-tillage and different cover crop alternatives to upland aerobic rice in tropical
114 environments (Nascente *et al.*, 2013). But, to the best of our knowledge, there is no
115 information available about rice production in Europe involving sprinkler irrigated aerobic
116 systems combined with different soil management regimes such as no-tillage and conservation
117 agriculture practices.

118 The aim of the present study was therefore to determine the short- and long-term
119 effects of sprinkler vs flood irrigation and no-tillage vs conventional tillage on soils' physical,
120 physicochemical, and biological properties, as well as on the yield components and
121 productivity in semi-arid Mediterranean conditions.

122 **2. Materials and Methods**

123 **2.1. Description of the study area**

124 A multi-year field experiment was conducted in 2011, 2012, and 2013 on a Hydragric
125 Anthrosol (FAO, 2006) with 16.9% clay, 35.9% silt, and 47.2% sand. The experimental field is
126 located in Extremadura, SW Spain (39°06' N; 5°40' W), with a Mediterranean climate (annual
127 rainfall < 480 mm), with hot, dry summers. This region has serious limitations in developing
128 sustainable agriculture as it has an aridity index of 0.49 (UNESCO, 1977).

129 **2.2. Field experiment**

130 Prior to beginning the study, the experimental area (1800 m²) was cropped with rice
131 (*Oryza sativa* L.) using the traditional management practices in the region (deep ploughing and
132 waterlogging), and a part of the field had already been devoted to direct seeding and sprinkler-

133 irrigated rice in the 7 years preceding the experiment. After harvesting the rice in November
134 2010, the field was divided into twelve plots of 140 m² (7×20 m) each, that were subjected to
135 the following four management regimes: (CTF) applying the techniques that are conventional
136 in the region, *i.e.*, tillage to 30 cm and flooding with continuous water flow; (CTS) conventional
137 tillage and sprinkler irrigation; (NTS) conservation agriculture techniques (no-tillage and
138 seeding by direct drilling) and sprinkler irrigation; and (NTS7) the same conservation
139 agriculture techniques (no-tillage and direct drilling) and sprinkler irrigation but where this
140 management regime had already been in use for 7 years in order to observe any long-term
141 changes. Each treatment was replicated thrice in a completely randomized design with two
142 protective buffer zones between adjacent plots (4.5×20 m). Flooded plots were separated with
143 ridges 35 cm high in order to avoid water losses.

144 **2.3. Crop culture**

145 Mouldboard ploughing was applied to 30 cm depth prior to sowing in the CTS and CTF
146 treatments. In the NTS and NTS7 treatments, each year after harvest, the crop residues were
147 left on the soil surface and the soils were left untilled. In contrast, all crop residues were
148 withdrawn from the CTS and CTF plots. The crop was sown in May, using Semeato trailed
149 pneumatic seed drills for the direct drilling of the NTS7 and NTS treatments, conventional
150 pneumatic seed drills for the CTS treatment, and a broadcast seed drill for the CTF treatment
151 after initial flooding. Rice was sown in 0.2 m rows in NTS7, NTS, and CTS. The sowing rate was
152 160 kg ha⁻¹ in all treatments. The seeds were of *Oryza sativa* L. var. *gladio*, a genotype
153 traditionally used in the Extremadura region. For NTS7, NTS, and CTS, a solid-set sprinkler
154 irrigation system was designed to cover the entire plot area. The irrigation water supplied in
155 the different treatments was monitored using a water flow-meter. The total amount of
156 irrigation water applied in 2011 was 7010 m³ ha⁻¹ for NTS7, NTS, and CTS, and 24 400 m³ ha⁻¹
157 for CTF. In 2012, it was 6705 m³ ha⁻¹ for NTS7, NTS, and CTS, and 34 290 m³ ha⁻¹ for CTF. And in

158 2013, it was 7800 m³ ha⁻¹ for NTS7, NTS, and CTS, and 32 235 m³ ha⁻¹ for CTF. The amount of
159 irrigation water applied for the sprinkler irrigated treatments was enough to keep soil
160 moisture over 70% of field capacity (Stevens *et al.*, 2012). For the flood irrigation, the aim was
161 to keep the water level constant. In all three years (2011, 2012, and 2013), composite fertilizer
162 9-18-27 (550 kg ha⁻¹) was applied in April as basal in all treatments, and N was applied in the
163 form of urea in two splits of 200 kg ha⁻¹ at tillering and 150 kg ha⁻¹ at the panicle initiation
164 stage. Pesticide applications were made using a backpack sprayer. Pre-emergence herbicides
165 were applied yearly – Pendimethalin in the sprinkler irrigated treatments (NTS7, NTS, and CTS)
166 and Oxadiazon in the flooded treatment (CTF). Penoxulam and Cyhalofop were used as post-
167 emergence herbicides in all treatments for first weed generations, and Bispyribac-Sodium for
168 late-emergence weeds. All other recommended culture practices for maximum grain yield
169 were followed.

170 **2.4. Soil sampling and analysis**

171 Soil samples from each plot were taken using a manual auger. The depths were: 0-20 cm
172 for physicochemical properties; 0-10 cm for enzymatic activities; and 0-10 and 10-30 cm for
173 aggregate stability. Three subsamples were taken randomly from each of the three replicate
174 plots at the beginning of the study (March 2011), and each year after harvest (October 2011,
175 2012, and 2013). In October 2011 and 2013, when the soil water content was near field
176 capacity (Bradford, 1986), the soil penetration resistance was measured in the field down to 45
177 cm depth using a hand penetrometer with 1 cm² conical tip. Texture was determined by
178 sedimentation using the Robinson pipette method (Soil Conservation Service, 1972) after
179 destruction of the organic carbon with H₂O₂ and chemical dispersion using Na₄P₂O₇. Total
180 organic carbon (TOC) content was determined by dichromate oxidation (Nelson & Sommers,
181 1996). Water soluble organic carbon (WSOC) was extracted with de-ionized water at a 3:1
182 water-to-soil ratio. Humic acids (HA) and fulvic substances (fulvic acids + humins, FA) were

183 extracted with a solution of 0.1 M Na₄P₂O₇ + NaOH at a 10:1 extractant-to-sample ratio, and, to
 184 precipitate humic acid, the supernatant was acidified to pH 2 with H₂SO₄. The WSOC and the
 185 TOC associated with each fraction of HA and FA were determined by dichromate oxidation and
 186 measurement of the absorbance at 590 nm (Sims & Haby, 1971). A combination electrode was
 187 used to measure electrical conductivity (EC) in a saturated soil extract and the pH in a 1:1 (w/v)
 188 soil/water mixture. Total nitrogen (N) was determined by the Kjeldahl method (Bremner &
 189 Mulvaney, 1982), and phosphorus (P) by the Olsen method (Olsen *et al.*, 1954). Aggregate
 190 stability (AS) was determined following the Sun method (Sun *et al.*, 1995) which uses a single
 191 0.250 mm sieve and an apparatus with a stroke length of 1.3 cm and frequency of 35
 192 cycles min⁻¹, and sodium hydroxide as dispersant. Table 1 gives some selected properties of
 193 the soil samples taken at the beginning of the study (March 2011).

194 **Table 1.** Soil physicochemical properties prior to the trial.

	TOC (g kg ⁻¹)	HA (g kg ⁻¹)	FA (g kg ⁻¹)	EC (μS cm ⁻¹)	pH	N (%)	P (mg kg ⁻¹)
NTS7	15.3	1.22	1.03	0.472	6.24	0.119	20.6
NTS	8.56	0.481	0.626	0.733	5.54	0.101	19.2
CTS	8.68	0.459	0.594	0.740	5.46	0.097	19.5
CTF	8.50	0.460	0.605	0.739	5.53	0.098	19.3

195 **TOC:** Total organic carbon; **HA:** Humic acid; **FA:** Fulvic acid; **EC:** Electrical conductivity; **N:** Total nitrogen; **P:**
 196 Phosphorus.

197 **2.5. Enzyme activities**

198 Dehydrogenase (DH) activity was determined by the method of Trevors (1984) as
 199 modified by García *et al.* (1993). An aliquot of 1 g of soil was incubated for 20 h at 20°C in
 200 darkness with 0.2 mL of 0.4% 2-(4-iodophenyl)-3-(4-nitrophenyl)-5-phenyl-2H-tetrazolium
 201 (INT) chloride as substrate. At the end of the incubation the idonitrotetrazolium formazan
 202 produced was extracted with 10 mL of methanol, and the absorbance measured at 490 nm.
 203 The DH and other activities were determined in triplicate. To assay the urease (UR) activity,
 204 2 mL of 0.1 M pH 7.0 phosphate buffer and 0.5 mL of 1.066 M urea were added to 0.5 g of soil

205 and incubated for 1.5 h at 30°C. The ammonia released in the hydrolytic reaction was
206 measured spectrophotometrically at 636 nm (Kandeler & Gerber, 1988; Nannipieri *et al.*,
207 1980). The activity of β -glucosidase (GLU) was determined by incubating 1 g of soil with 4 mL
208 of 25 mM 4-nitrophenyl- β -D-glucopyranoside in 0.1 M modified universal buffer (MUB) pH 6.0
209 (Tabatabai, 1982). For the determination of phosphatase (PHO) activity, 4 mL of 25 mM 4-
210 nitrophenyl phosphate MUB pH 11 was added to 1 g of soil (Tabatabai & Bremner, 1969). For
211 the determination of arylsulfatase (ARS) activity, 4 mL of 5 mM 4-nitrophenyl sulfate in 0.5 M
212 acetate buffer pH 5.8 was added to 1 g of soil (Tabatabai & Bremner, 1970). The soil samples
213 were incubated for 1 h at 37°C, then cooled to 2°C for 15 min to halt the reaction. The p-
214 nitrophenol produced in the enzymatic reactions was determined at 400 nm, 398 nm, and
215 410 nm for GLU, PHO, and ARS, respectively. Blank assays without soil or substrate were
216 performed at the same time as the controls.

217 **2.6. Agronomic parameters**

218 All agronomic parameters were determined from a 2 m² sampling area within each plot
219 (three plots per treatment), and all production parameters were standardized to a moisture
220 content of 14% fresh weight. The total number of grains per panicle and the ripening ratio
221 were determined by direct counting on 20 representative panicles. The 1000 grain weight was
222 determined by direct weighing of 1000 grains counted by an electronic seed counter
223 (Swantech-SC2). Grain production was determined as the direct weight of all filled grains per
224 panicle collected in the trial area. Biomass production was determined by the direct weight of
225 the aerial part of the plant after grain removal. The harvest index (HI) was taken to be the ratio
226 between the yield and the total biomass (yield + straw yield), and water productivity was taken
227 to be the ratio between yield and the applied amount of irrigation water.

228 **2.7. Statistical analyses**

229 Statistical analyses were performed using the IBM SPSS Statistics 22.0 program package.
230 The data were checked for normality and homoskedasticity. A one-way ANOVA was used to
231 analyse soil and agronomic properties, the Duncan test to determine significant parameter
232 differences between treatments and years, and the Pearson correlation coefficient to study
233 possible correlations between different parameters. Differences were considered statistically
234 significant at a p-value of less than 0.05.

235 **3. Results and Discussion**

236 **3.1. The soils' physicochemical properties**

237 The soils' physicochemical parameters at 0-20 cm depth for the years 2011, 2012, and
238 2013 are listed in Table 2. All the TOC values were low, as is usual for agricultural soils in
239 Mediterranean environments (López-Garrido *et al.*, 2012). The highest values were for NTS7 in
240 all three years (15.6, 15.5, and 16.2 g kg⁻¹), reflecting a major effect of long-term no-tillage
241 practices on this parameter. Although increases in TOC have been observed in rice cultivation
242 under flood irrigation (Hao *et al.*, 2013), in the present case there were no significant
243 variations of the TOC content in the NTS treatment over the three years studied (Table 2).
244 Indeed, the short-term effects of no-tillage practices on TOC are complex, with widely varying
245 results being reported that depend on such factors as climate, crop residue, and the crop
246 management regime itself (Muñoz *et al.*, 2007). On the other hand, the CTS and CTF
247 treatments showed significant declines (by 23.9% and 13.4%, respectively) in TOC content for
248 2013 compared to 2011. The irrigation method is an important factor that may help explain
249 the differences in TOC losses between the CTS and CTF treatments. In this sense, it is
250 important to note that organic matter mineralization is related to soil moisture balance, in
251 particular, to the air-water ratio. Aerobic microbial activity in soil rises with increasing
252 moisture content up to limits at which the amount of water reduces oxygen availability. The

253 anaerobic conditions caused by the flooding in the CTF treatment may have resulted in a
 254 slower rate of organic matter decomposition than in the CTS treatment, in which the
 255 conditions for faster organic matter mineralization were more favourable.

256 **Table 2.** Effect of different management regimes on soil physicochemical properties.

	TOC (g kg ⁻¹)	HA (g kg ⁻¹)	FA (g kg ⁻¹)	EC (μS cm ⁻¹)	pH	N (%)	P (mg kg ⁻¹)
2011							
NTS7	15.6cA	1.60cB	1.03cA	1.51dC	5.84cA	0.143bB	16.4bA
NTS	7.93aA	0.833bB	0.532aA	0.906bB	5.91cB	0.081aA	11.5aA
CTS	9.23abC	0.604aB	0.676bB	0.698aA	5.63bA	0.074aA	9.26aA
CTF	9.92bB	0.627aB	0.634bA	1.26cB	5.15aA	0.086aB	11.1aA
2012							
NTS7	15.5cA	1.33bA	0.993cA	0.490aA	5.92dA	0.106bA	19.0bB
NTS	6.80aA	0.514aA	0.579aB	0.857bA	5.92cB	0.072aA	18.0aB
CTS	8.22bB	0.515aA	0.685bB	1.09cB	5.84bB	0.078aA	21.6cC
CTF	8.25bA	0.481aA	0.661bA	1.27dB	5.21aB	0.079aA	26.0dC
2013							
NTS7	16.2dA	1.76dC	0.998dA	1.26cB	5.84cA	0.173dC	20.8cC
NTS	7.67bA	1.13cC	0.764cC	1.38dC	5.22aA	0.104bB	21.9dC
CTS	7.03aA	0.908aC	0.586aA	0.753aA	6.24dC	0.094aB	12.5aB
CTF	8.60cA	0.984bC	0.639bA	1.08bA	5.33bC	0.114cB	17.2bB
Y	**	***	**	***	*	***	***
T	***	***	***	***	***	***	***
Y*T	***	***	***	***	***	***	***

257 **TOC:** Total organic carbon; **HA:** Humic acid; **FA:** Fulvic acid; **EC:** Electrical conductivity; **N:** Total nitrogen; **P:**
 258 Phosphorus. ANOVA factors are **Y:** Year; **T:** Treatment; **Y*T:** Interaction Year*Treatment; *, ** and *** significant at
 259 α levels of 0.05, 0.01 and 0.001, respectively; **NS:** not significant. Different letters show significant differences
 260 (p<0.05) between treatments in the same year (lower case letters) and between years within the same treatment
 261 (upper case letters).

262 The greatest FA content throughout the trial corresponded to NTS7. There were no
 263 significant changes in this parameter between years for the NTS7 or CTF treatments (Table 2).
 264 It is important to recall that the management techniques applied in these two treatments did
 265 not differ from those applied prior to the experiment, reflecting a long-term stabilization. But
 266 the NTS and CTS treatments showed contrasting trends: while the FA content in NTS increased

267 by 43.6% in 2013 relative to 2011 due to the annual input of organic matter, the FA content in
268 CTS fell by 13.3% over the same period.

269 With respect to the HA content, all the treatments presented significant decreases in
270 2012 relative to 2011 (Table 2), and reached their highest values in 2013. That this pattern is
271 independent of the type of management and irrigation technique points to the influence of
272 external factors. In particular, high rainfall was recorded in autumn 2012, with very wet
273 months of September, October, and November months (AEMET, 2012). There was some 40
274 mm of rainfall prior to the 2012 soil sampling, which may have slowed down organic matter
275 humification in all the treatments. Then, the more favourable weather conditions in 2013
276 generated the major rise in HA content due to the cumulative humification of organic matter
277 left over from 2012 as well as that of 2013.

278 The EC values were highly variable, reflecting the great spatial variability of this
279 parameter in field conditions. In 2013, after three years of the trial, the lowest value of this
280 parameter corresponded to the CTS soil and the highest to the two no-tillage treatments (NTS7
281 and NTS). The TOC accumulation in the no-tillage treatments and the organic matter depletion
282 in CTS may help to explain these contrasting trends. In the literature, there is no consistent
283 pattern in the effects of no-tillage practices on EC. While in a long-term experiment (13 years)
284 Dalal (1989) observed EC values that were lower in soils under no-tillage than under
285 conventional tillage, Pérez-Brandán *et al.* (2012) observed the contrary, although neither of
286 those works studied the influence of the irrigation system.

287 The treatments greatly influenced the pH levels. There was acidification of the NTS7 and
288 NTS soils in 2013 relative to 2012. This was probably because of a build-up of organic residues
289 on the surface as was noted by Limousin & Tessier (2007) in studying topsoil acidification in
290 no-tillage maize and wheat plots. On the other hand, the pH of the CTS soil rose (by 10.8%)
291 from 5.63 in 2011 to 6.25 in 2013. The CTF soil showed a similar trend, but with a much smaller

292 increase in pH (by 3.50%). This behaviour in the two conventionally tilled treatments is related
293 to the changes in TOC content discussed above, with the rises in pH being explicable by the
294 losses of organic matter in these treatments (Table 2).

295 The total nitrogen (N) was significantly correlated with the TOC content ($r=0.769$;
296 $p<0.001$), with, in all three years, the highest values corresponding to the NTS7 treatment and
297 the lowest to CTS. This effect of long-term no-tillage on soil nitrogen is coherent with the
298 observations of other workers (Heenan *et al.*, 2004; Pérez-Brandán *et al.*, 2012) for different
299 crops and environmental conditions, with, in all cases, there being significant increases in N
300 under no-tillage relative to conventional tillage systems.

301 In 2011, NTS7 had the highest P levels, while there were no significant differences
302 between the rest of the treatments. In 2012, the highest levels corresponded to the
303 conventional tillage soils (CTS and CTF). This trend was reversed in 2013, however, when the
304 highest values corresponded to the no-tillage treatments (NTS7 and NTS). While a study
305 involving sprinkler irrigated corn (Essington & Howard, 2000) found no significant differences
306 in the P content of soils under conventional and no-tillage practices, the present results are
307 coherent with those of a study of agricultural soils in southern Spain (Bravo *et al.*, 2006) in
308 that, in the mid to long term, no-tillage practices may enhance the P content in Mediterranean
309 soils.

310 Aggregate stability (AS), expressed as the stable fraction percentage (Figure 1), is a key
311 factor in soil fertility (Hernanz *et al.*, 2002). It has been found to be significantly correlated with
312 TOC ($r=0.935$; $p<0.001$), FA ($r=0.773$; $p<0.001$), and HA ($r=0.891$; $p<0.001$) for the 0-10 cm
313 depth, and with TOC ($r=0.656$; $p<0.01$), FA ($r=0.907$; $p<0.001$), and HA ($r=0.867$; $p<0.001$) for
314 the 10-30 cm depth. These results highlight the importance of organic matter and humic
315 substances in stabilizing soil structure. For this reason, the highest values of AS corresponded
316 to NTS7 for all three years and for both depths. In the NTS treatment, there was a significant

317 increase in AS in 2013 in the 0-10 depth, and even more importantly in the 0-30 depth. The
 318 CTS and CTF treatments also showed increases in AS in 2013 reflecting the greater HA content,
 319 but these increases were not comparable to those of the no-tillage treatments.

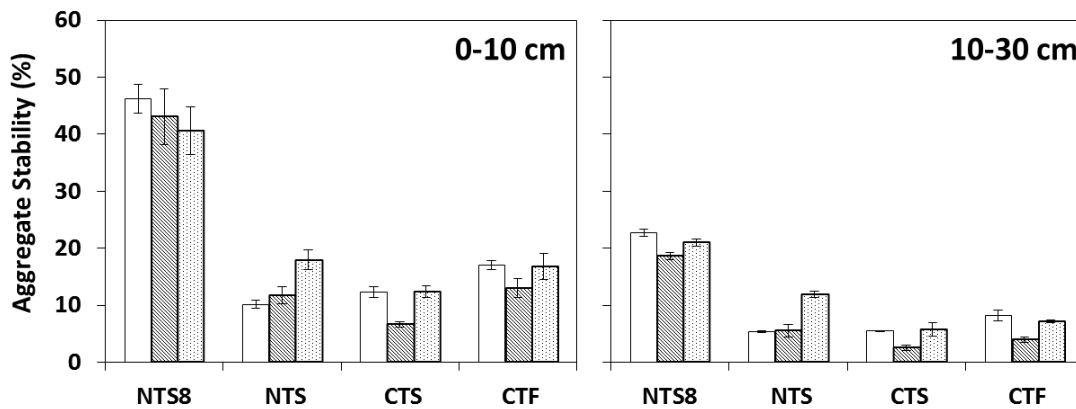


Figure 1. Effect of different management regimes on soil aggregate stability in 2011 (□), 2012 (▨), and 2013 (▩).

320 Figure 2 shows the penetration resistance results for 2011 and 2013. One observes that,
 321 for both years, there was greater compaction in the upper layers of soil for the no-tillage
 322 treatments (NTS7 and NTS), and that the compaction of the conventionally tilled treatments
 323 (CTS and CTF) increased with depth to surpass that of the no-tillage soils, until reaching a
 324 maximum at 20-25 cm. Beyond this depth, the CTS and CTF values decreased down to levels
 325 similar to those for NTS7 and NTS. This pattern suggests that the conventional tillage
 326 generated a plough pan, and highlights the importance of taking depth into account when
 327 considering variations in penetration resistance (Afzalnia & Zabihi, 2014). The pattern of the
 328 two no-tillage curves is also coherent with the finding of a study on rainfed wheat cropping
 329 with a no-tillage regime (Ferrerias *et al.*, 2000) that penetration resistance increased faster with
 330 depth in the top layers of the soil than in deeper layers.

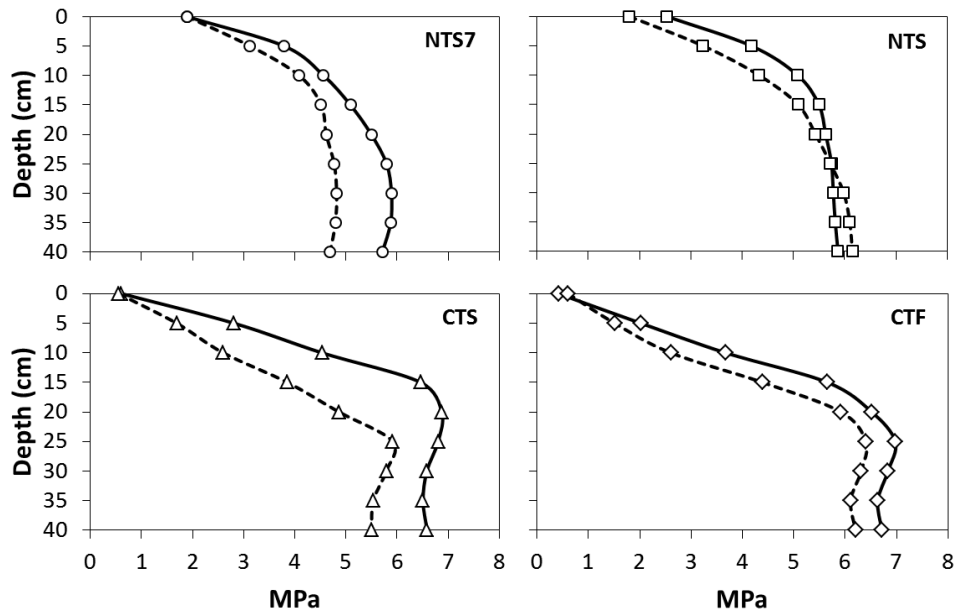


Figure 2. Effect of different management regimes on soil penetration resistance. Treatments are: NTS7 (○), NTS (□), CTS (▲) and CTF (◇). Years are: 2011 (—) and 2013 (---).

331 3.2. Enzymatic activities

332 Dehydrogenase activity (DH) is a good indicator of the microbial redox system, and is
 333 also regarded as a good measure of soil microbial activity (Moreno *et al.*, 2008). It was found
 334 to be greatest in all years in the NTS7 soil (Table 3), with there being no significant differences
 335 between the rest of the treatments in 2012 and 2013. The TOC content of NTS7 may have
 336 influenced this behaviour as it showed a significant correlation with DH activity ($r=0.709$;
 337 $p<0.01$).

338 A soil's β -glucosidase activity (GLU) plays an important part in hydrolytic processes
 339 during organic matter degradation (Pandey *et al.*, 2014). The highest values corresponded to
 340 the NTS7 treatment (Table 3), and they were significantly correlated with the TOC content
 341 ($r=0.808$; $p<0.01$), indicating that the addition of crop residues is a good strategy with which to
 342 increase GLU in soils.

343 Urease activity (UR) is of major importance since it is related to the nitrogen cycle, and
 344 also because of the extensive use of urea as a nitrogen fertilizer. A significant decrease in UR
 345 was observed in 2013 relative to 2012 for all treatments except NTS7. The increase in N

346 contents in 2013 as a consequence of applications of urea to the crop may have had an
 347 inhibitory effect on UR, in coherence with the findings of Pandey *et al.* (2014) in aerobic rice
 348 fields under no-tillage and conventional tillage management.

349 **Table 3.** Effect of different management regimes on soil enzyme activities.

	DH ($\mu\text{g INTF g}^{-1} \text{h}^{-1}$)	GLU ($\mu\text{mol pNP g}^{-1} \text{h}^{-1}$)	UR ($\mu\text{mol NH}_4^+ \text{g}^{-1} \text{h}^{-1}$)	PHO ($\mu\text{mol pNP g}^{-1} \text{h}^{-1}$)	ARS ($\mu\text{mol pNP g}^{-1} \text{h}^{-1}$)
2011					
NTS7	0.948cA	0.522bA	3.09bA	1.71bA	0.150bA
NTS	0.854bB	0.119aA	1.23aB	1.30aA	0.072aA
CTS	0.750bB	0.156aA	1.22aB	1.40aA	0.088abA
CTF	0.645aB	0.196aA	1.49aB	1.23aA	0.096abA
2012					
NTS7	0.781bA	0.846bAB	3.27cA	1.84bA	0.138bA
NTS	0.455aA	0.288aB	1.49bB	1.31aA	0.041aA
CTS	0.466aA	0.270aB	1.02aAB	1.41aA	0.060aA
CTF	0.516aA	0.287aAB	1.54bB	1.31aA	0.049aA
2013					
NTS7	0.884bA	1.11bB	3.13bA	1.62bA	0.207bA
NTS	0.527aA	0.298aB	0.766aA	1.07aA	0.068aA
CTS	0.689aB	0.340aB	0.809aA	1.11aA	0.080aA
CTF	0.481aA	0.372aB	0.782aA	1.21aA	0.109aA
Y	***	***	***	***	**
T	***	***	***	***	***
Y*T	**	*	*	NS	NS

350 **ARS:** Arylsulfatase activity; **GLU:** β -glucosidase activity; **DH:** Dehydrogenase activity; **PHO:** Phosphatase activity; **UR:**
 351 Urease activity. ANOVA factors are **Y:** Year; **T:** Treatment; **Y*T:** Interaction Year*Treatment; *, ** and *** significant
 352 at α levels of 0.05, 0.01 and 0.001, respectively; **NS:** not significant. Different letters show significant differences
 353 ($p < 0.05$) between treatments in the same year (lower case letters) and between years within the same treatment
 354 (upper case letters).

355 An apparently similar trend was observed for the phosphatase activity (PHO), but the
 356 decreases in 2013 for all treatments did not reach statistical significance. But the NTS7 values
 357 were indeed significantly higher than the rest, reflecting that there is more substrate available

358 for microbial activity in this treatment than in other treatments with less time of no-tillage
359 management (Shi *et al.*, 2013).

360 Arylsulfatase activity (AR) is one of the commonest enzymatic activities in soils. Besides
361 satisfying microbial and plant needs for sulfate ion, it takes part in the processes of degrading
362 xenobiotic compounds in soils (Kertesz *et al.*, 1994). For this reason, it is important to try to
363 gain some insight into how AR varies in response to different management techniques.
364 Coherent with the findings of Lin (2007) in comparing AR between flooded and upland aerobic
365 rice crops, we found no significant differences between the flooded and sprinkler irrigated
366 treatments (Table 3). For the NTS7 soil, there was an increase in AR in 2013. This may have
367 been related to the pH value (6.1 from 0 to 10 cm) which is in the range for this enzyme's
368 greatest activity. This is supported by the significant correlation found between AR and pH
369 ($r=0.623$; $p<0.01$).

370 **3.3. Yield, yield attributes, and water productivity**

371 Table 4 presents the results for the yield, yield attributes, and water productivity. It is
372 important to note that reductions in yield due to weed competition have been recognized as
373 one of the main threats for sprinkler irrigated rice since the first field experiments of Westcott
374 & Vines (1986) and McCauley (1990). Given the abnormal circumstances of 2011 when there
375 was a very high level of weed infestation over the entire area of the Guadiana River lowlands,
376 our discussion will mainly focus on the 2012 and 2013 results.

377 For these two years, NTS7 had the greatest number of grains per panicle (GP), although
378 the difference was statistically significant only in 2012. There were no significant differences
379 between the other treatments, irrespective of the irrigation and tillage methods applied. The
380 values of GP were significantly correlated with TOC ($r=0.469$; $p<0.01$) and HA ($r=0.664$;
381 $p<0.001$). This latter correlation was also found by Saha *et al.* (2013) in studying the influence

382 of a soil's humic acids on rice plant growth, and may explain the greater value of GP in the
 383 NTS7 treatment.

384 **Table 4.** Effect of different management regimes on yield and yield components of rice.

	GP	RR (%)	1000W (g)	Y (kg ha ⁻¹)	HI	WP (g L ⁻¹)
2011						
NTS7	101bA	82.4aA	22.5aA	4 621bA	0.49abA	0.66cA
NTS	70.3aA	79.0aA	21.5aA	2 519aA	0.44aA	0.36bA
CTS	82.5aA	83.3aC	22.8aB	2 328aA	0.50bB	0.33bA
CTF	65.4aA	85.2aA	21.9aA	4 550bA	0.45aA	0.19aA
2012						
NTS7	121cA	74.8bA	22.0bA	9 805eC	0.53dA	1.46dC
NTS	94.8abB	69.7bA	21.4abA	4 844bB	0.46bA	0.72cB
CTS	82.7aA	60.6aA	20.4aA	3 590aB	0.35aA	0.53bB
CTF	88.4aB	83.8cA	25.4cB	6 556cB	0.49cB	0.19aA
2013						
NTS7	109aA	84.4bcA	23.8cA	7 397cB	0.60aA	0.94cB
NTS	99.5aB	81.8bB	23.0bcB	8 229dC	0.61aB	1.05dC
CTS	106aB	75.9aB	21.9aB	4 784aC	0.57aB	0.61bB
CTF	95.3aB	89.9dA	24.9dB	8 926eC	0.68bC	0.27aB
Y	***	***	**	***	***	***
T	**	***	***	***	**	***
Y*T	NS	**	***	***	***	***

GP: Grains per panicle; **RR:** Ripening ratio; **1000W:** 1000 grain weight; **Y:** Yield; **HI:** Harvest index; **WP:** Water productivity. ANOVA factors are **Y:** Year; **T:** Treatment; **Y*T:** Interaction Year*Treatment; ******, and ******* significant at α levels of 0.01, and 0.001, respectively; **NS:** not significant. Different letters show significant differences ($p < 0.05$) between treatments in the same year (lower case letters) and between years within the same treatment (upper case letters).

385 The ripening ratio (RR) (Table 4) was significantly greater ($p < 0.05$) under flooded
 386 conditions (CTF) in both years. Any reduction in the amount of water available may turn into a
 387 situation of moisture stress for the rice plant, and hence affect grain ripening in the aerobic
 388 treatments (NTS7, NTS, and CTS) relative to flooded conditions (Wei *et al.*, 2011). The lowest
 389 ripening ratio in both 2012 and 2013 corresponded to CTS, indicating that, under aerobic

390 conditions, no-tillage gives the soil a greater moisture retention capacity, therefore reducing
391 the frequency and intensity of moisture stress situations in the NTS7 and NTS treatments
392 relative to CTS.

393 The pattern of the 1000 grain weight results for both years (2012 and 2013) was similar
394 to that of RR, with significantly ($p < 0.05$) greater weights in the flooded treatment (CTF) than in
395 the aerobic treatments, and with CTS being the treatment presenting the lowest weight.

396 In 2011, the yields for each treatment were lower than in the other two years. There
397 was no significant difference between the NTS7 (4621 kg ha^{-1}) and the CTF (4550 kg ha^{-1})
398 yields, indicating that long-term implementation of no-tillage and sprinkler irrigation
399 management may achieve yields that are similar to those obtained under traditional
400 management techniques, but, in the present case, with the benefit of 71.2% of water savings.

401 In 2012, the NTS7 treatment gave the highest yield (9805 kg ha^{-1}), 49.5% higher than
402 that of the CTF treatment (6556 kg ha^{-1}) while using 80.4% less water. It is important to note
403 that the CTF treatment was on a paddy soil with relatively high percolation losses. This NTS7
404 yield was also greater than the yields recorded by Stevens *et al.* (2012) for different rice
405 varieties under sprinkler irrigation and conventional tillage. The NTS and CTS yields were lower
406 than that of CTF. In the short term, no-tillage may lead to significant yield reductions in maize
407 and other crops (Linden *et al.*, 2000), so that similar reductions could be expected in aerobic
408 rice too.

409 In 2013, there was a significant reduction in the NTS7 yield, while the yields of the other
410 three treatments each attained their greatest values of the three years. This may have been a
411 sign of negative effects of long-term rice monoculture, in which case crop rotation would be
412 necessary in order both to avoid the selection of herbicide-resistant weeds and to improve the
413 soil's fertility. Nevertheless, the NTS7 yield of 7397 kg ha^{-1} was still very close to the Spanish
414 average under flooded conditions (7600 kg ha^{-1}). Although the CTF treatment gave the greatest

415 yield (8926 kg ha^{-1}), it required more than four times the amount of water that was applied in
416 the sprinkler irrigated treatments. The NTS yield was far greater than in the previous two
417 years, reaching 8229 kg ha^{-1} . These results show that no-tillage combined with sprinkler
418 irrigation has the potential of providing good yields in the mid-term, and that this potential
419 may be maintained in the long term. Although the CTS yield was also greater in 2013
420 (4784 kg ha^{-1}), this was significantly ($p < 0.05$) less than the yields obtained with the other
421 treatments, and even less than the Regional average, indicating that the combination of
422 sprinkler irrigation and conventional tillage is poorly suited to the environmental conditions of
423 the study. A correlation analysis for the 2012 and 2013 data showed yield to be significantly
424 correlated with TOC ($r = 0.542$; $p < 0.01$), FA ($r = 0.583$; $p < 0.01$), HA ($r = 0.622$; $p < 0.01$), and N
425 ($r = 0.476$; $p < 0.01$), indicating the important influence of the quantity and quality of soil organic
426 matter on rice crop productivity.

427 The harvest index (HI) values for 2011 were low due to the low yields achieved by all
428 four treatments. For 2012, the values were within the mean value ranges reported by Bueno &
429 Lafarge (2009) for different rice varieties and genotypes. For 2013, there was a significant
430 ($p < 0.05$) increase in the HI values for all the treatments (Table 4). While there were no
431 significant differences between the sprinkler irrigated treatments, the CTF value was the
432 highest of the entire experiment, close to the limits indicative of a high risk of lodging.

433 The management regimes significantly influenced the water productivity (WP) values in
434 the three years (Table 4), but the effects were different from year to year as determined by the
435 significant treatment \times year interaction ($p < 0.001$; Table 4). One of the main goals of rice-based
436 cropping systems worldwide is to increase water productivity (WP). In all three years, the CTF
437 plot had the lowest WP. The mid-term effect of implementing no-tillage for sprinkler irrigated
438 treatments was clearly noticeable in 2012, with NTS7 having the greatest WP (1.46 g L^{-1}),
439 followed by NTS (0.72 g L^{-1}), and then CTS (0.53 g L^{-1}). In the third year of study, the greatest

440 WP corresponded to NTS (1.05 g L⁻¹), followed by NTS7 (0.94 g L⁻¹), whereas CTF showed the
441 lowest value (0.27 g L⁻¹). These results were consistent with findings of previous studies by
442 Lampayan et al. (2015) who reported that water productivities in aerobic rice managements
443 (0.86-1.24 g L⁻¹) were higher than anaerobic rice (0.50-0.63 g L⁻¹). There were significant
444 correlations ($p < 0.01$) of WP with the soils' TOC ($r = 0.615$), HA ($r = 0.632$), FA ($r = 0.730$) contents.
445 These results suggest that conservation agriculture practices such as no-tillage, which increase
446 the soil's organic matter content, may have a boosting effect on rice crops' WP under
447 Mediterranean environmental conditions. With regard to other studies carried out in
448 Mediterranean environments, the 2012 NTS7 WP was greater than the values reported by
449 Spanu *et al.* (1997) for conventionally tilled and sprinkler irrigated rice, and the 2013 NTS7 and
450 NTS WP values were in the range of the most productive varieties tested by Guiducci *et al.*
451 (1998) when comparing the adaptation of different rice varieties to sprinkler irrigation (0.79-
452 1.15 g L⁻¹), although neither of these studies applied conservation agriculture techniques. The
453 2013 CTS WP value was below the range reported by Guiducci *et al.* (1998) for the least
454 productive varieties they tested (0.72-0.78 g L⁻¹). An explanation might be the TOC depletion in
455 this soil during the course of the present study.

456 **4. Conclusions**

457 This study has shown that aerobic rice production combined with no-tillage practices
458 may induce important transformations in the soil, leading to major improvements in its
459 physical, physicochemical, and biological properties. This is especially so after long-term
460 implementation of the strategy. Although, in the short-term, sprinkler irrigation gave lower
461 yields, in the mid-term, the combination of this irrigation method with no-tillage practices gave
462 yields similar to those observed in the conventional tillage plus flood irrigation treatment, but
463 with water savings that averaged 75% of the total amount of water used in the flood irrigation.
464 Moreover, these yields, water savings, and soil improvements are sustainable in the long-term.

465 The greatest water productivity values corresponded to the no-tillage, aerobic rice systems. In
466 sum, the mid- and long-term implementation of no-tillage combined with sprinkler irrigation
467 may be considered to be a productive and sustainable management system for rice farming
468 under semi-arid Mediterranean conditions.

469 **Acknowledgements**

470 This research was supported by Spain's Ministry of Science and Innovation and Ministry
471 of the Economy and Competitiveness (Projects AGL2010-21421-C02-02 and AGL2013-48446-
472 C3-2-R). D. Becerra and D. Peña were recipients of a grant from the Consejería of Economía,
473 Comercio e Innovación of the Government of Extremadura.

474 **References**

- 475 AEMET (Spanish Meteorological Agency). 2012. Informe meteorológico y agrofenológico de
476 otoño de 2012. Madrid.
- 477 Afzalnia, S.; Zabihi, J. 2014. Soil compaction variation during corn growing season under
478 conservation tillage. *Soil and Tillage Research*, 137: 1-6. Doi: 10.1016/j.still.2013.11.003.
- 479 Barker, R.; Dawe, D.; Tuong, T.P.; Bhuiyan, S.I.; Guerra, L.C. 1999. The outlook for water
480 resources in the year 2020: challenges for research on water management in rice
481 production, in: FAO (Ed.), *Assesment and orientation towards the 21st century*, El Cairo,
482 Egypt. pp. 96-109.
- 483 Bezdicek, D.F.; Beaver, T.; Granatstein, D. 2003. Subsoil ridge tillage and lime effects on soil
484 microbial activity, soil pH, erosion, and wheat and pea yield in the Pacific Northwest,
485 USA. *Soil and Tillage Research* 74: 55-63. Doi: 10.1016/S0167-1987(03)00091-6.
- 486 Bradford, J.M. 1986. Penetrability, in: A. Klute (Ed.), *Methods of soilanalysis. Part 1. Physical*
487 *and Mineralogical methods*, pp. 468–471.

488 Bravo, C.; Torrent, J.; Giráldez, J.V.; González, P.; Ordóñez, R. 2006. Long-term effect of tillage
489 on phosphorus forms and sorption in a Vertisol of southern Spain. *European Journal of*
490 *Agronomy*, 25(3): 264-269. Doi: 10.1016/j.eja.2006.06.003.

491 Bremner, J.M.; Mulvaney, C.S. 1982. Nitrogen-Total. *Methods of Soil Analysis, Soil Science*
492 *Society of America, Special Publication 9*: 595-624.

493 Bueno, C.S.; Lafarge, T. 2009. Higher crop performance of rice hybrids than of elite inbreds in
494 the tropics. *Field Crops Research*, 112:229-237. Doi: 10.1016/j.fcr.2009.03.006.

495 CGIAR Science Council, 2006. IRRI's upland rice research follow-up review to the 6th IRRI EPMR.
496 Rome, Italy, Science Council Secretariat.

497 Dalal, R.C. 1989. Long-Term Effects of No-Tillage, Crop Residue, and Nitrogen Application on
498 Properties of a Vertisol. *Soil Science Society of America Journal*, 53: 1511-1515. Doi:
499 10.2136/sssaj1989.03615995005300050035x.

500 Devkota, K.P.; Lamers, J.P.; Manschadi, A.M.; Devkota, M.; McDonald, A.J.; Vlek, P.L. 2015.
501 Comparative advantages of conservation agriculture based rice–wheat rotation systems
502 under water and salt dynamics typical for the irrigated arid drylands in Central Asia.
503 *European Journal of Agronomy*, 62: 98-109. Doi: 10.1016/j.eja.2014.10.002.

504 Essington, M.; Howard, D. 2000. Phosphorus availability and speciation in long-term no-till and
505 disk-till soil. *Soil Science* 165: 144-152. Doi: 10.1097/00010694-200002000-00005.

506 Facchi, A.; Gharsallah, O.; Chiaradia, E.A.; Bischetti, G.B.; Gandolfi, C. 2013. Monitoring and
507 modelling evapotranspiration in flooded and aerobic rice fields. *Procedia Environmental*
508 *Sciences*, 19: 794-803. Doi: 10.1016/j.proenv.2013.06.088.

509 FAO. 2006. Guidelines for soil description. Food and Agriculture Organization of the United
510 Nations, Rome, Italy.

511 Ferreras, L.A.; Costa, J.L.; Garcia, F.O.; Pecorari, C. 2000. Effect of no-tillage on some soil
512 physical properties of a structural degraded etrocalcic aleudoll of the southern “Pampa”
513 of Argentina. *Soil and Tillage Research* 54: 31-39. Doi: 10.1016/S0167-1987(99)00102-6.

514 García, C.; Hernandez, T.; Costa, C.; Ceccanti, B.; Masciandaro, G.; Ciardi, C. 1993. A study of
515 biochemical parameters of composted and fresh municipal wastes. *Bioresources*
516 *Technology*, 44: 17–23. Doi: 10.1016/0960-8524(93)90202-M.

517 Guiducci, M.; Bonciarelli, U.; Benincasa, P.; Rosati, E. 1998. Potenzialittá produttiva e
518 adattamento di varietà di riso irrigate per aspersione in Umbria. *L'Informatore Agrario*,
519 16: 31-35.

520 Hao, Q.; Cheng, B.; Jiang, C. 2013. Long-term tillage effects on soil organic carbon and
521 dissolved organic carbon in a purple paddy soil of Southwest China. *Acta Ecologica*
522 *Sinica*, 33: 260-265. Doi: 10.1016/j.chnaes.2013.07.005.

523 Heenan, D.P.; Chan, K.Y.; Knight, P.G. 2004. Long term impact of rotation, tillage and stubble
524 management on the loss of soil organic carbon and nitrogen from a Chromic Luvisol. *Soil*
525 *and Tillage Research* 76: 59-68. Doi: 10.1016/j.still.2003.08.005.

526 Hernanz, J.; López, R.; Navarrete, L.; Sánchez-Girón. 2002. Long-term effects of tillage systems
527 and rotations on soil structural stability and organic carbon stratification in semiarid
528 central Spain. *Soil and Tillage Research*, 66(2): 129-141. Doi: 10.1016/S0167-
529 1987(02)00021-1.

530 Kandeler, E.; Gerber, H. 1988. Short-term assay of soil urease activity using colorimetric
531 determination of ammonium. *Biology and Fertility of Soils*, 6: 68–72. Doi:
532 10.1007/BF00257924.

533 Kertesz, M.A.; Kölbener, P.; Stockinger, H.; Beil, S.; Cook, A.M. 1994. Desulfonation of linear
534 alkylbenzenesulfonate surfactants and related compounds by bacteria. *Applied*
535 *Environmental Microbiology*, 60(7): 2296-2303.

536 Khumar, V.; Ladha, J.K. 2011. Direct seeding of rice: recent developments and future research
537 needs. *Advances in agronomy*, 111: 297-396. Doi: 10.1016/B978-0-12-387689-8.00001-
538 1.

539 Lin, S. 2007. A comparative study of paddy soils and upland soil enzyme activity and organic
540 matter in the FT - IR characteristics. Plant Nutrition Department, Yangzhou University,
541 Yangzhou, China. pp. 55.

542 Lanfranchi, M.; Giannetto, C.; De Pascale, A. 2014. Economic implications of climate change for
543 agricultural productivity. *WSEAS Transactions on Environment and Developmet* 10: 233-
544 241.

545 Limousin, G.; Tessier, D. 2007. Effects of no-tillage on chemical gradients and topsoil
546 acidification. *Soil and Tillage Research*, 92: 167-174. Doi: 10.1016/j.still.2006.02.003.

547 Linden, D.R.; Clapp, C.E.; Dowdy, R.H. 2000. Long-term corn grain and stover yields as a
548 function of tillage and residue removal in east central Minnesota. *Soil and Tillage*
549 *Research*, 56(3-4): 167-174. Doi: 10.1016/S0167-1987(00)00139-2.

550 López-Garrido, R.; Deurer, M.; Madejón, E.; Murillo, J.M.; Moreno, F. 2012. Tillage influence on
551 biophysical soil properties: The example of a long-term tillage experiment under
552 Mediterranean rainfed conditions in South Spain. *Soil and Tillage Research* 118: 52-60.
553 Doi: 10.1016/j.still.2011.10.013.

554 López-Garrido, R.; Madejón, E.; Moreno, F.; Murillo, J.M. 2014. Conservation tillage influence
555 on carbon dynamics under Mediterranean conditions. *Pedosphere*, 24: 65-75. Doi:
556 10.1016/S1002-0160(13)60081-8.

557 MAGRAMA (Spanish Ministry of Agriculture). 2013. Cultivos herbáceos e industriales: el arroz.
558 Madrid, Spain.

559 McCauley, G.N. 1990. Sprinkler vs. flood irrigation in traditional rice production regions of
560 southeast Texas. *Agronomy Journal* 82(4): 677-683. Doi:
561 10.2134/agronj1990.00021962008200040006x.

562 Moreno, B.; Vivas, A.; Nogales, R.; Macci, C.; Masciandaro, G.; Benítez, E. 2008. Restoring
563 biochemical activity and bacterial diversity in a trichloroethylene-contaminated soil: the
564 reclamation effect of vermicomposted olive wastes. *Environmental Science and*

565 *Pollution Research*, 16(3): 253-264. Doi: 10.1007/s11356-008-0035-y.

566 Moreno-Jiménez, E.; Meharg, A.; Smolders, E.; Manzano, R.; Becerra, D.; Sánchez-Llerena, J.;
567 Albarrán, A.; López-Piñeiro, A. 2014. Sprinkler irrigation of rice fields reduces grain
568 Arsenic but enhances Cadmium. *Science of the Total Environment* 485-486: 468-473. Doi:
569 10.1016/j.scitotenv.2014.03.106.

570 Mukhopadhyay, M.; Datta, J.K.; Garai, T.P. 2013. Steps toward alternative farming system in
571 rice. *European Journal of Agronomy*, 51: 18-24. Doi: 10.1016/j.eja.2013.06.005.

572 Muñoz, A.; López-Piñeiro, A.; Ramírez, M. 2007. Soil quality attributes of conservation
573 management regimes in a semi-arid region of south western Spain. *Soil and Tillage*
574 *Research*, 95: 255-265. Doi: 10.1016/j.still.2007.01.009.

575 Nannipieri, P.; Ascher, J.; Ceccherini, M.T.; Landi, L.; Pietramellara, G.; Renella, G.
576 2003. Microbial diversity and soil functions. *European Journal of Soil Sciences*. 54: 655–
577 670. Doi: 10.1046/j.1351-0754.2003.0556.x.

578 Nascente, A.S.; Crusciol, C.A.C.; Cobucci, T. 2013. The no-tillage system and cover crops
579 alternatives to increase upland rice yields. *European Journal of Agronomy*, 45: 124-131.
580 Doi: 10.1016/j.eja.2012.09.004.

581 Nelson, D.W.; Sommers, L.E. 1996. Total carbon, organic carbon and organic matter. In: Sparks
582 DL (ed.) *Methods of soil analysis, Part 3*. Soil Sciences Society Book Ser. 5 SSSA. Madison,
583 pp 961–1010.

584 Olsen, S.R.; Cole, C.V.; Watanabe, F.S. 1954. Estimation of available phosphorus in soils by
585 extraction with sodium bicarbonate. United States Government Publication Office,
586 Washington, USA.

587 Pandey, D.; Agrawal, M; Singh, J. 2014. Effects of conventional tillage and no-tillage
588 permutations on extracellular soil enzyme activities and microbial biomass under rice
589 cultivation. *Soil and Tillage Research*, 136: 51-60. Doi: 10.1016/j.still.2013.09.013.

590 Pérez-Brandán, C.; Arzeno, J.L.; Huidobro, J.; Grümberg, B.; Conforto, C.; Hilton, S.; Bending,
591 G.D.; Mireles, J.M.; Vargas-Gil, S. 2012. Long-term effect of tillage systems on soil
592 microbiological, chemical and physical parameters and the incidence of charcoal rot by
593 *Macrophomina phaseolina* (Tassi) Goid in soybean. *Crop protection*, 40: 73-82. Doi:
594 10.1016/j.cropro.2012.04.018.

595 Porter, J.R. 2005. Rising temperatures are likely to reduce crop yields. *Nature*, 436 (7048): 174.
596 Doi: 10.1038/436174b.

597 Rajakumar, D.; Subramanian, E.; Ramesh, T.; Maragatham, N.; Martin, G.J.; Thiagarajan, G.
598 2009. Striding towards aerobic rice cultivation. *Agricultural Reviews*, 30: 213-218. Doi:
599 10.5958/0976-0741.2014.

600 Russo, S.; Nardi, S. 1996. Effetti della coltura asciutta sulla qualità del granello. *L'Informatore*
601 *Agrario*, 12: 25-32.

602 Saha, R.; Saieed, M.; Chowdhury, M. 2013. Growth and yield of rice (*Oryza sativa* L.) as
603 influenced by humic acids and poultry manure. *Universal Journal of Plant Sciences*, 1 (3):
604 78-84. Doi: 10.13189/ujps.2013.010304.

605 Shi, Y.; Lalande, R.; Hamel, C.; Ziadi, N.; Gagnon, B.; Hu, Z. 2013. Seasonal variation of microbial
606 biomass, activity, and community structure in soil under different tillage and phosphorus
607 management practices. *Biology and Fertility of Soils*, 49(7): 803-818. Doi:
608 10.1007/s00374-013-0773-y.

609 Sims, J.R.; Haby, V.A. 1971. Simplified colorimetric determination of soil organic matter. *Soil*
610 *Science*, 112: 137-141. Doi: 10.1097/00010694-197108000-00007.

611 Soil Conservation Service. 1972. Soil survey laboratory methods and procedures for collecting
612 soil samples. Washington, USA.

613 Spanu, A.; Pruneddu, G.; Porreca, P. 1997. Prove su riso irrigato per aspersione. *L'Informatore*
614 *Agrario*, 13: 51-54.

615 Stevens, G.; Vories, E.; Heiser, J.; Rhine, M. 2012. Experimentation on cultivation of rice
616 irrigated with a center pivot system. In: S. Lee (Ed.), *Irrigation systems and practices in*
617 *challenging environments, InTech*. pp. 233-254.

618 Sun, H.; Larney, F.J.; Bullock, M.S. 1995. Soil amendments and water-stable aggregation of a
619 desurfaced dark brown chernozem. *Canadian Journal of Soil Sciences*, 75: 319-325. Doi:
620 10.4141/cjss95-046

621 Tabatabai, M.A. 1982. Soil enzymes. In: Page, A.L., Miller, E.M., Keeney, D.R.
622 (Eds.), *Methods of Soil Analyses. Part 2. Chemical and Microbiological*
623 *Properties*, ASA, Madison, pp. 903–947.

624 Tabatabai, M.A.; Bremner, J.M. 1969. Use of p-nitrophenyl phosphate for assay of soil
625 phosphatase activity. *Soil Biology and Biochemistry*, 1: 301–307. Doi: 10.1016/0038-
626 0717(69)90012-1.

627 Tabatabai, M.A.; Bremner, J.M. 1970. Arylsulphatase activity of soils. *Soil Sciences Society of*
628 *America Proceedings*, 34: 225–229. Doi: 10.2136/sssaj1970.03615995003400020016x.

629 Trevors; J.T. 1984. Dehydrogenase activity in soil: A comparison between the INT
630 and TTC assay. *Soil Biology and Biochemistry*. 16: 673–674. Doi: 10.1016/0038-
631 0717(84)90090-7.

632 UNESCO. 1977. Map of the world distribution of Arid Regions. MAB. UNESCO, Paris, Technical
633 Notes 7.

634 Van Tran, D. 2002. Concerns about rice production practices. 20th Session of International Rice
635 Commission, Bangkok, Thailand, 23-26 July.

636 Wei, F.; Tao, H.; Lin, S.; Bouman, B.A.; Zhang, L.; Wang, P.; Dittert, K. 2011. Rate and duration
637 of grain filling of aerobic rice HD297 and their influence on grain yield under different
638 growing conditions. *Science Asia*, 37:98-114. Doi: 10.2306/scienceasia1513-
639 1874.2011.37.098.

640 Westcott, M.P.; Vines, K.W. 1986. A comparison of sprinkler and flood irrigation for rice.

